# Characterization of Potential Adverse Health Effects Associated with Consuming Fish from 

Ellison Creek Reservoir<br>"Lone Star Lake"

Morris County, Texas

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Department of State Health Services
Division for Regulatory Services
Policy, Standards, and Quality Assurance Unit
Seafood and Aquatic Life Group
Austin, Texas

## TABLE OF CONTENTS

LIST OF FIGURES ..... 1
LIST OF TABLES ..... 1
LIST OF ACRONYMS ..... 3
SUMMARY ..... 5
INTRODUCTION ..... 6
History of the Ellison Creek Reservoir Fish Consumption Advisory ..... 6
The TMDL Program at the TCEQ and the Relationship between the TMDL Program and Consumption Advisories or Possession Bans Issued by the DSHS ..... 7
Description of Ellison Creek Reservoir ..... 8
Population of Morris County Surrounding Ellison Creek Reservoir ..... 8
Subsistence Fishing at Ellison Creek Reservoir ..... 9
METHODS ..... 9
Fish Sampling, Preparation, and Analysis ..... 9
Fish Sampling Methods and Description of the Ellison Creek Reservoir 2014 Sample Set ..... 9
Fish Age Estimation ..... 11
Analytical Laboratory Information. ..... 11
Details of Some Analyses with Explanatory Notes ..... 12
Calculation of Dioxin Toxicity Equivalence (TEQ) ..... 14
Derivation and Application of Health-Based Assessment Comparison Values for Systemic (Noncarcinogenic) Effects ( $\mathrm{HAC}_{\text {nonca }}$ ) of Consumed Chemical Contaminants ..... 15
Derivation and Application of Health-Based Assessment Comparison Values for Application to the Carcinogenic Effects ( $\mathrm{HAC}_{\text {ca }}$ ) of Consumed Chemical Contaminants ..... 18
Children’s Health Considerations ..... 19
Data Analysis and Statistical Methods ..... 19
RESULTS ..... 20
Inorganic Contaminants ..... 21
Organic Contaminants ..... 21
DISCUSSION ..... 26
Risk Characterization ..... 26
Characterization of Systemic (Noncarcinogenic) Health Effects from Consumption of Fish from Ellison Creek Reservoir ..... 27
Characterization of Theoretical Lifetime Excess Cancer Risk from Consumption of Fish from Ellison Creek Reservoir. ..... 28
Characterization of Calculated Cumulative Systemic (Noncarcinogenic) Health Effects and of Cumulative Excess Lifetime Cancer Risk from Consumption of Fish from Ellison Creek Reservoir. 30
CONCLUSIONS ..... 31
RECOMMENDATIONS ..... 32
PUBLIC HEALTH ACTION PLAN ..... 34
TABLES ..... 50
LITERATURE CITED ..... 79

## LIST OF FIGURES

FIGURE 1. ELLISON CREEK RESERVOIR MAP ..... 36
FIGURE 2. THE RELATIONSHIP BETWEEN PCB CONCENTRATION AND TOTAL LENGTH FOR FISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 37
FIGURE 3. THE RELATIONSHIP BETWEEN PCB CONCENTRATION AND PERCENT LIPIDS FOR FISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 38
FIGURE 4. LENGTH AT AGE FOR CHANNEL CATFISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 39
FIGURE 5. THE RELATIONSHIP BETWEEN PCB CONCENTRATION AND TOTAL LENGTH FOR CHANNEL CATFISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 40
FIGURE 6. THE RELATIONSHIP BETWEEN PCB CONCENTRATION AND AGE FOR CHANNEL CATFISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014 ..... 41
FIGURE 7. LENGTH AT AGE FOR CRAPPIE COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 42
FIGURE 8. LENGTH AT AGE FOR FLATHEAD CATFISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 43
FIGURE 9. LENGTH AT AGE FOR LARGEMOUTH BASS COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 44
FIGURE 10. THE RELATIONSHIP BETWEEN PCB CONCENTRATION AND PERCENT LIPID FOR LARGEMOUTH BASS COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014.45
FIGURE 11. THE RELATIONSHIP BETWEEN PCDD/PCDF CONCENTRATION AND TOTAL LENGTH FOR FISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 46
FIGURE 12. THE RELATIONSHIP BETWEEN PCDD/PCDF CONCENTRATION AND PERCENT LIPID FOR FISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 47
FIGURE 13. THE RELATIONSHIP BETWEEN PCDD/PCDF CONCENTRATION AND TOTAL LENGTH FOR CHANNEL CATFISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 48
FIGURE 14. THE RELATIONSHIP BETWEEN PCB CONCENTRATION AND PERCENT LIPID FOR CHANNEL CATFISH COLLECTED FROM ELLISON CREEK RESERVOIR, TEXAS, 2014. ..... 49
LIST OF TABLES
TABLE 1. FISH SAMPLES COLLECTED FROM ELLISON CREEK RESERVOIR 2014. SAMPLE NUMBER, SPECIES, TOTAL LENGTH, AND WEIGHT RECORDED FOR EACH SAMPLE. ..... 50
TABLE 2.2. CADMIUM (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 54
TABLE 2.4. LEAD (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 56
TABLE 2.5. SELENIUM (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SPECIES, 2014 ..... 57
TABLE 2.6. ZINC (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 58
TABLE 2.7. MERCURY (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SAMPLE SITE, 2014 ..... 59
TABLE 2.8. MERCURY (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SAMPLE SITE, 2014 ..... 60
TABLE 3. PESTICIDES (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 62
TABLE 4.1. PCBS (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SAMPLE SITE, 2014 ..... 63
TABLE 4.2. PCBS (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SAMPLE SITE, 2014 ..... 64
TABLE 4.3. PCBS (MG/KG) IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 65
TABLE 5.1. PCDDS/PCDFS TOXICITY EQUIVALENT (TEQ) CONCENTRATIONS (PG/G) IN FISH COLLECTED FROM THE ELLISON CREEK RESERVOIR BY SAMPLE SITE, 2014. ..... 66
TABLE 5.2. PCDDS/PCDFS TOXICITY EQUIVALENT (TEQ) CONCENTRATIONS (PG/G) IN FISH COLLECTED FROM THE ELLISON CREEK RESERVOIR BY SAMPLE SITE, 2014. ..... 67
TABLE 5.3. PCDDS/PCDFS TOXICITY EQUIVALENT (TEQ) CONCENTRATIONS (PG/G) IN FISH COLLECTED FROM THE ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 68
TABLE 6.0. VOLATILE ORGANIC COMPOUNDS (MG/KG) IN FISH COLLECTED FROM THE ELLISON CREEK RESERVOIR BY SPECIES, 2014. ..... 69
TABLE 7. HAZARD QUOTIENTS (HQS) FOR MERCURY IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR IN 2014. ..... 70
TABLE 8.1. HAZARD QUOTIENTS (HQS) AND HAZARD INDICES (HIS) FOR PCBS AND/OR PCDDS/PCDFS IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR IN 2014 ..... 71
TABLE 8.2. HAZARD QUOTIENTS (HQS) AND HAZARD INDICES (HIS) FOR PCBS AND/OR PCDDS/PCDFS IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR IN 2014 ..... 72
TABLE 8.3. HAZARD QUOTIENTS (HQS) AND HAZARD INDICES (HIS) FOR PCBS AND/OR PCDDS/PCDFS IN FISH COLLECTED FROM ELLISON CREEK RESERVOIR IN 2014 ..... 73
TABLE 9.1. CALCULATED THEORETICAL LIFETIME EXCESS CUMULATIVE CANCER RISK FROM CONSUMING FISH COLLECTED IN 2014 FROM ELLISON CREEK RESERVOIR CONTAINING CARCINOGENS AND SUGGESTED CONSUMPTION RATE ..... 74
TABLE 9.2. CALCULATED THEORETICAL LIFETIME EXCESS CUMULATIVE CANCER RISK FROM CONSUMING FISH COLLECTED IN 2014 FROM ELLISON CREEK RESERVOIR CONTAINING CARCINOGENS AND SUGGESTED CONSUMPTION RATE ..... 75
TABLE 9.3. CALCULATED THEORETICAL LIFETIME EXCESS CUMULATIVE CANCER RISK FROM CONSUMING FISH COLLECTED IN 2014 FROM ELLISON CREEK RESERVOIR CONTAINING CARCINOGENS AND SUGGESTED CONSUMPTION RATE ..... 76
TABLE 9.4. CALCULATED THEORETICAL LIFETIME EXCESS CUMULATIVE CANCER RISK FROM CONSUMING FISH COLLECTED IN 2014 FROM ELLISON CREEK RESERVOIR CONTAINING CARCINOGENS AND SUGGESTED CONSUMPTION RATE ..... 77
TABLE 10. SALG RECOMMENDED FISH CONSUMPTION ADVICE FOR ELLISON CREEK RESERVOIR, 2014. ..... 78

## LIST OF ACRONYMS

| ARL | Acceptable Lifetime Risk Level |
| :---: | :---: |
| ATSDR | Agency for Toxic Substances and Disease Registry |
| BDL | Below Detection Limit |
| BMD | Benchmark Dose |
| BMDL | Benchmark Dose (Lower Confidence Limit) |
| ca | Cancer |
| CDC | Centers for Disease Control |
| CPF | Cancer Potency Factor |
| CSF | Cancer Slope Factor |
| DDD | Dichlorodiphenyldichloroethane |
| DDE | Dichlorodiphenyldichloroethylene |
| DDT | Dichlorodiphenyltrichloroethane |
| dL | Deciliter |
| DSHS | Department of State Health Services |
| g | Gram |
| GC | Gas Chromatograph |
| GERG | Geochemical and Environmental Research Group |
| GSMFC | Gulf States Marine Fisheries Commission |
| HAC | Health Assessment Comparison |
| HCH | Hexachlorocyclohexane |
| HI | Hazard Index |
| HQ | Hazard Quotient |
| in | Inches |
| IH | Interstate Highway |
| IRIS | Integrated Risk Information System |
| kg | Kilogram |
| lb | Pound |
| LOAEL | Lowest Observed Adverse Effects Level |
| mcg | Microgram |
| mg | Milligram |
| mm | Millimeter |
| MRL | Minimal Risk Level |
| MS | Mass spectrometer |
| n | Sample Size |
| ND | Not Detected |
| NOAA | National Oceanic and Atmospheric Administration |
| NOAEL | No Observed Adverse Effects Level |
| nonca | Noncancer |
| p | Statistical Significance in a Hypothesis Test |
| PCB | Polychlorinated Biphenyl |
| PCDD | Polychlorinated Dibenzo-p-Dioxin |
| PCDF | Polychlorinated Dibenzofuran |

## LIST OF ACRONYMS CONT.

| pg | picogram |
| :--- | :--- |
| $r$ | Correlation Coefficient |
| $r^{2}$ | Coefficient of Determination |
| RfD | Reference Dose |
| RL | Reporting Limit |
| SALG | Seafood and Aquatic Life Group |
| SOP | Standard Operating Procedure |
| SSD | Seafood Safety Division |
| SVOC | Semivolatile Organic Compound |
| TCEQ | Texas Commission on Environmental Quality |
| TDH | Texas Department of Health |
| TEF | Toxicity Equivalence Factor |
| TEQ | Toxicity Equivalence |
| TL | Total Length |
| TMDL | Total Maximum Daily Load |
| TNRCC | Texas Natural Resources Conservation Commission |
| TPWD | Texas Parks and Wildlife Department |
| UL | Intake Level |
| USEPA | United States Environmental Protection Agency |
| VOC | Volatile Organic Compound |
| $\bar{X}$ | Mean |

## SUMMARY

A survey of Ellison Creek Reservoir Lone Star, Texas in 2005 indicated that polychlorinated biphenyl concentrations in fish exceeded Texas Department of Health guidelines for protection of human health. Since 2005, the Texas Department of State Health Services has recommended that people do not eat fish from Ellison Creek Reservoir.

In 2014, the Texas Department of State Health Services performed this study to investigate any potential change in fish tissue contamination in Ellison Creek Reservoir. The present study examined fish from Ellison Creek Reservoir for the presence and concentrations of environmental toxicants that, if eaten, potentially could affect human health negatively. The study also addresses the public health implications of consuming fish from Ellison Creek Reservoir and suggests actions to reduce potential adverse health outcomes.

Results of the 2014 survey indicate that polychlorinated biphenyl and dioxin concentrations in channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, sunfish Spp., and white bass continue to exceed Texas Department of State Health Services guidelines for protection of human health.

## Conclusions

- Confidence in the conclusions for many species of fish is limited by the small sample size. Sampling a small number of fish (i.e., individual species of fish or all fish species combined) decreases the confidence of mean contaminant concentrations for the fish population thus adding uncertainty to the conclusions.
- Regular or long-term consumption of channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, sunfish Spp., and white bass may result in adverse systemic (noncarcinogenic) health effects. Therefore, consumption of these species of fish from Ellison Creek Reservoir poses an apparent risk to human health.
- Regular or long-term consumption of channel catfish, common carp, and hybrid striped bass may increase the likelihood of carcinogenic health risks. Therefore, consumption of these species of fish from Ellison Creek Reservoir poses an apparent risk to human health.


## Recommendations

- People should not consume common carp and hybrid striped bass from Ellison Creek Reservoir (Table 10).
- Women of childbearing age (Women and girls under 50) including pregnant women,
women who may become pregnant, and women who are nursing infants and children less than 12 years of age, or who weigh less than 75 pounds should not consume channel catfish, common carp, flathead catfish, hybrid striped bass, spotted gar, sunfish Spp., and white bass from Ellison Creek Reservoir.
- Women of childbearing age (Women and girls under 50) including pregnant women, women who may become pregnant, and women who are nursing infants and children less than 12 years of age, or who weigh less than 75 pounds may consume up to one four-ounce meal per month of largemouth bass from Ellison Creek Reservoir.
- Women past childbearing age (Women 50 and older) and males 12 and older may consume up to one eight-ounce meal per month of channel catfish, flathead catfish, spotted gar, or sunfish Spp. from Ellison Creek Reservoir.
- Women past childbearing age (Women 50 and older) and males 12 and older may consume up to two eight-ounce meals per month of largemouth bass or white bass from Ellison Creek Reservoir.
- The DSHS advise TPWD to continue not stocking hybrid striped bass (i.e., Palmetto bass) in Ellison Creek Reservoir because hybrid striped bass bioaccumulate significant concentrations of PCBs and PCDDs/PCDFs that pose apparent hazards to public health. The TPWD discontinued stocking of hybrid striped bass following the issuance of the Ellison Creek Reservoir fish consumption advisory in 2005.


## INTRODUCTION

This document summarizes the results of a survey of Ellison Creek Reservoir (ECR) also known as Lone Star Lake conducted in 2014 by the Texas Department of State Health Services (DSHS) Seafood and Aquatic Life Group (SALG). ${ }^{\text {a }}$ The SALG did this study to investigate any potential change in fish tissue contamination in ECR. The present study examined fish from ECR for the presence and concentrations of environmental toxicants that, if eaten, potentially could affect human health negatively. The report addresses the public health implications of consuming fish from ECR and suggests actions to reduce potential adverse health outcomes.

## History of the Ellison Creek Reservoir Fish Consumption Advisory

The Texas Commission on Environmental Quality (TCEQ) surveyed water and sediments from Ellison Creek Reservoir for chemical contaminants between March 1, 1998 and February 28, 2003. Following evaluation of this data, the TCEQ listed Ellison Creek Reservoir in the 2004 Texas Water Quality Inventory and on 303(d) List. In the 2004 Texas Water Quality Inventory, the TCEQ expressed "concern for toxicity in sediment to aquatic organisms in the southeast

[^0]part of the reservoir near the Lone Star facility due to ... elevated levels of metal contaminants in sediment". ${ }^{1}$

In December 2003, the SALG examined analytical data for nine fish fillets and five whole fish samples collected between June 2002 and July 2003 by TCEQ regional personnel. Laboratory analysis of these samples revealed several to contain lead and/or PCBs. Four of nine fish fillet samples contained PCBs at concentrations ranging from 0.150 to $0.320 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.090 \mathrm{mg} / \mathrm{kg}$. Three of five whole fish samples contained PCBs ranging from not detected (ND) to $0.210 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.120 \mathrm{mg} / \mathrm{kg}$. The mean concentration of PCBs in combined fish tissues collected from Ellison Creek Reservoir in 2002 and 2003 were approximately twice the DSHS guideline for protection of human health from exposure to PCBs (PCB Health Assessment Comparison value $0.047 \mathrm{mg} / \mathrm{kg}$ ).

In the 2002-2003 TCEQ data set, two fish fillet samples contained lead at levels below the laboratory's practical quantitation limit (PQL) for lead ( $4 \mathrm{mg} / \mathrm{kg}$ ). Whole fish samples did not contain demonstrable lead. The SALG risk assessors were unable to assess the significance of lead in 2002-2003 fish collected from Ellison Creek Reservoir: because they could not determine whether lead concentrations were listed as dry-weight or wet-weight concentrations; because lead concentrations were below the PQL; and, because only two fish fillet samples contained lead, while no whole fish were found to contain measurable lead.

From the TCEQ 2002-2003 data for Ellison Creek Reservoir, the SALG risk assessors suggested to TCEQ that further investigation of fish from the reservoir to characterize adequately human health risks associated with consuming contaminants in fish from Ellison Creek Reservoir was necessary. Subsequent to that review, the TCEQ and the SALG determined that Ellison Creek Reservoir was a candidate for examination under the Statewide Fish Tissue Monitoring Program. In 2005, SALG staff performed a survey of Ellison Creek Reservoir with funding provided by the TCEQ through the Statewide Fish Tissue Monitoring Program.

The 2005 survey revealed the presence of PCBs at concentrations exceeding DSHS health-based guidelines in fish from Ellison Creek Reservoir. ${ }^{2}$ The DSHS issued Fish and Shellfish Consumption Advisory 29 (ADV-29) on November 28, 2005 advising people not to consume fish from Ellison Creek Reservoir. ${ }^{3}$

## The TMDL Program at the TCEQ and the Relationship between the TMDL Program and Consumption Advisories or Possession Bans Issued by the DSHS

The TCEQ enforces federal and state laws that promote judicious use of water bodies under state jurisdiction and protects state-controlled water bodies from pollution. Pursuant to the federal Clean Water Act, Section 303(d), ${ }^{4}$ all states must establish a "total maximum daily load" (TMDL) for each pollutant contributing to the impairment of a water body for one or more designated uses. A TMDL is the maximum amount of a pollutant that a body of water can assimilate and still meet water quality standards. ${ }^{5}$ TMDLs incorporate margins of safety to ensure the usability of the water body for all designated purposes. States, territories, and tribes
define the uses for a specific water body (e.g., drinking water, contact recreation, aquatic life support) along with the scientific criteria designated to support each specified use.

Fish consumption is a recognized use for many waters. A water body is impaired if fish from that water body contain contaminants that make those fish unfit for human consumption or if consumption of those contaminants potentially could harm human health. Although a water body and its aquatic life may clear toxicants over time with removal of the source(s), it is often necessary to institute some type of remediation such as those implemented by the TCEQ. Thus, whenever the DSHS issues a fish consumption advisory or prohibits possession of environmentally contaminated fish, the TCEQ places the water body in its current Texas Integrated Report of Surface Water Quality formerly called the Texas Water Quality Inventory and $303(\mathrm{~d})$ List. ${ }^{6}$ The TCEQ is responsible for confirming the impairment and, if necessary, the TMDL program, then prepares a TMDL for each contaminant present at concentrations that, if consumed, would be capable of negatively affecting human health. After approval of the TMDL, the stakeholders in the watershed prepare an Implementation Plan for each contaminant. These plans are designed to facilitate the rehabilitation of the water body over time. Successful remediation should result in return of the water body to conditions compatible with all stated uses, including consumption of fish from the water body. When the DSHS lifts a consumption advisory or possession ban, people may once again keep and consume fish from the water body. If fish in a water body are contaminated, one of the several items on an Implementation Plan for a water body on a state's 303(d) list consists of the periodic reassessment of contaminant levels in resident fish.

## Description of Ellison Creek Reservoir

Ellison Creek Reservoir (also known as Lone Star Lake) is a 1,516-acre impoundment of Ellison Creek. ${ }^{7}$ The reservoir is located in Morris County within the City of Lone Star, Texas. The reservoir is owned and operated by U.S. Steel Tubular Products, Inc. (formerly owned by Lone Star Steel). Ellison Creek (also known as Bruton's Creek) is a tributary of Big Cypress Creek and part of the Cypress River Basin. It has a watershed that spans 37 square miles, a shoreline of 14 miles, and a maximum depth of 40 feet with moderate water clarity. Reservoir structural habitat is sparse and is comprised of inundated timber, brush, riprap, creek channels, and boat docks. Aquatic vegetation is also present in the reservoir. Predominant fish species include largemouth bass, hybrid striped bass, spotted bass, channel catfish, white bass, crappie, redbreast sunfish, bluegill, and redear sunfish. Texas Parks and Wildlife Department (TPWD) statewide harvest regulations govern management of fish species taken from Ellison Creek Reservoir. The reservoir has two public boat ramps and one privately operated ramp. ${ }^{8}$

## Population of Morris County Surrounding Ellison Creek Reservoir

Ellison Creek Reservoir is located in Morris County in rural Northeast Texas. The United States Census Bureau estimated the 2014 Morris County population at 12,743 people. ${ }^{9}$ Longview, Texas positioned approximately 30 miles south of Ellison Creek Reservoir is the closest metropolitan area (population $\geq 20,000$ people) in Northeast Texas. ${ }^{10}$

## Subsistence Fishing at Ellison Creek Reservoir

The USEPA suggests that, along with ethnic characteristics and cultural practices of an area's population, the poverty rate could contribute to any determination of the rate of subsistence fishing in an area. ${ }^{11}$ The USEPA and the DSHS find it is important to consider subsistence fishing to occur at any water body because subsistence fishers (as well as recreational anglers and certain tribal and ethnic groups) usually consume more locally caught fish than the general population. These groups sometimes harvest fish or shellfish from the same water body over many years to supplement caloric and protein intake. People, who routinely eat fish from chemically contaminated bodies of water or those who eat large quantities of fish from the same waters, could increase their risk of adverse health effects. The USEPA suggests that states assume that at least $10 \%$ of licensed fishers in any area are subsistence fishers. Subsistence fishing, while not explicitly documented by the DSHS, likely occurs in Texas. The DSHS assumes the rate of subsistence fishing to be similar to that estimated by the USEPA.

## METHODS

## Fish Sampling, Preparation, and Analysis

The DSHS SALG collects and analyzes edible fish from the state's public waters to evaluate potential risks to the health of people consuming contaminated fish or shellfish. Fish tissue sampling follows standard operating procedures from the DSHS Seafood and Aquatic Life Group Survey Team Standard Operating Procedures and Quality Control/Assurance Manual. ${ }^{12}$ The SALG bases its sampling and analysis protocols, in part, on procedures recommended by the USEPA's Guidance for Assessing Chemical Contaminant Data for Use in Fish Advisories, Volume $1 .{ }^{13}$ Advice and direction are also received from the Fish Sampling Advisory Subcommittee of the legislatively mandated State of Texas Toxic Substances Coordinating Committee. ${ }^{14}$ Samples usually represent species, trophic levels, and legal-sized specimens available for consumption from a water body. When practical, the DSHS collects samples from two or more sites within a water body to better characterize geographical distributions of contaminants.

## Fish Sampling Methods and Description of the Ellison Creek Reservoir 2014 Sample Set

In April-May 2014, the SALG staff collected 108 fish samples from Ellison Creek Reservoir. Risk assessors used data from these fish to assess the potential for adverse human health outcomes from consuming fish from this reservoir.

The SALG selected four sample sites to provide spatial coverage of the study area (Figure 1): Site 1 ECR at the dam; Site 2 ECR near American Electric Power intake; Site 3 ECR at Barnes Creek; and, Site 4 ECR near Upper Reservoir. Species collected represent distinct ecological groups (i.e., predators and bottom-dwellers) that have some potential to bio-accumulate chemical contaminants, have a wide geographic distribution, are of local recreational fishing value, and/or that anglers and their families commonly consume. The 108 fish collected from

Ellison Creek Reservoir represent all species targeted for collection from this water body (Table 1). The list below contains the number of each target species, listed in descending order collected for this study: channel catfish (35); largemouth bass (30); common carp (13); flathead catfish (8); crappie (7); sunfish species (4); spotted gar (4); bowfin (3); hybrid striped bass (2); and white bass (2).

The survey team set gill nets at sample sites 1-4 in late afternoon (Figure 1); fished the sites overnight, and collected samples from the nets early the following morning. The gill nets were set at locations to maximize available cover and habitat at each sample site. During collection, to keep specimens from different sample sites separated, the team placed samples from each site into mesh bags labeled with the site number. The survey team immediately stored retrieved samples on wet ice in large coolers to ensure interim preservation. Survey team members returned to the reservoir any live fish culled from the catch and properly disposed of samples found dead in the gill nets.

The survey team set trap nets at sample sites 1-4 to target crappie samples (Figure 1); fished the sites overnight, and collected samples from the nets early the following morning. The trap nets were set at locations to maximize available cover and habitat at each sample site. During collection, to keep crappie samples from different sample sites separated, the team placed samples from each site into mesh bags labeled with the site number. The survey team immediately stored retrieved samples on wet ice in large coolers to ensure interim preservation. Survey team members returned to the reservoir any live fish culled from the catch.

The SALG utilized a boat-mounted electrofisher to collect fish from samples sites 1-4. The SALG staff conducted electrofishing activities during daylight hours using pulsed direct current (Smith Root 7.5 GPP/ 5.0 GPP electrofishing system settings: 6.0-8.0 amps, 60 pulses per second [pps], high range, 500 volts, $60-100 \%$ duty cycle and 1.0-2.0 amps, 15 pps , high range, 500 volts, $80-$ $100 \%$ duty cycle) to stun fish that crossed the electric field in the water in front of the boat. Staff used dip nets over the bow of the boat to retrieve stunned fish, netting only fish preselected as target samples. Staff immediately stored retrieved samples on wet ice in large coolers to enhance tissue preservation.

The survey team utilized juglines (a fishing line with a three-way swivel, single circle hook, and bottom weight tied to a free-floating device) to catch flathead catfish. The SALG staff baited lines with live sunfish. The survey team targeted habitat likely to hold flathead catfish.

The SALG staff processed fish onsite at ECR. Staff weighed each sample to the nearest gram (g) on an electronic scale and measured total length (TL; tip of nose to tip of tail fin) to the nearest millimeter (mm; Table 1). All TL measurements were converted to inches for use in this report. After weighing and measuring a fish, staff used a cutting board covered with aluminum foil and a fillet knife to prepare two skin-off fillets from each fish. The foil was changed and knife cleaned with distilled water after each sample was processed. The SALG staff wrapped fillet(s) in two layers of fresh aluminum foil, placed in an unused, clean, pre-labeled plastic freezer bag,
and stored on wet ice in an insulated chest until further processing. The SALG staff transported tissue samples on wet ice to their Austin, Texas headquarters, where the samples were stored temporarily at $-5^{\circ}$ Fahrenheit ( $-20^{\circ}$ Celsius) in a locked freezer. The freezer key is accessible only to authorized SALG staff members to ensure chain of custody while samples are in the possession of agency staff. The SALG delivered the frozen fish tissue samples to the Geochemical and Environmental Research Group (GERG) Laboratory, Texas A\&M University, College Station, Texas, for contaminant analysis.

## Fish Age Estimation

The SALG staff removed sagittal otoliths from black and white crappie, channel catfish, flathead catfish, hybrid striped bass, largemouth bass, sunfish species, and white bass samples for age estimation. The SALG staff followed otolith extraction procedures recommended by the Gulf States Marine Fisheries Commission (GSMFC) and unpublished procedures recommended by the Texas Parks and Wildlife Department (TPWD). ${ }^{15,16}$ Staff performed all otolith extractions on each fish sample after the preparation of the two skin-off fillets for chemical contaminant analysis. Following extraction, staff placed otoliths in an individually labeled coin envelope and then in a plastic freezer bag to transport to their Austin, Texas headquarters. Staff processed otoliths and estimated ages according to procedures recommended by the GSMFC and TPWD. ${ }^{15,16}$

## Analytical Laboratory Information

The GERG personnel documented receipt of the 108 ECR samples and recorded the condition of each sample along with its DSHS identification number. Using established USEPA methods, the GERG laboratory analyzed fish fillets from ECR for inorganic and organic contaminants commonly identified in polluted environmental media. Analyses included seven metals (arsenic, cadmium, copper, lead, total mercury, selenium, and zinc), 123 semivolatile organic compounds (SVOCs), 70 volatile organic compounds (VOCs), 34 pesticides, 209 PCB congeners, ${ }^{\mathrm{b}, 17}$ and 17 polychlorinated dibenzo-p-dioxins and/or dibenzofurans (PCDDs/PCDFs) congeners. The laboratory analyzed all 108 samples for mercury and PCBs. A subset of 100 of the original 108 samples was analyzed for the following contaminant groups: metals and PCDDs/PCDFs, ${ }^{18}$ and a subset of 16 of the original 108 samples was analyzed for the following contaminant groups: pesticides, SVOCs, and VOCs. The SALG risk assessors selected the subset of samples based on target species and size class selection procedures outlined in SALG standard operating procedures (SOPs). In addition to SALG SOPs, if available, the SALG risk assessors use TPWD creel surveys to determine the species of fish most frequently harvested from the body of water being evaluated and choose large specimens of the selected species of fish. The SALG risk

[^1]assessors choose large fish to assess conservatively contaminant exposure when evaluating small sample sizes.

## Details of Some Analyses with Explanatory Notes

## Arsenic

The GERG laboratory analyzed 100 fish samples for total (inorganic arsenic + organic arsenic $=$ total arsenic) arsenic. Although the proportions of each form of arsenic may differ among fish species, under different water conditions, and, perhaps, with other variables, the scientific literature suggests that well over $90 \%$ of arsenic in fish is likely organic arsenic - a form of arsenic that is virtually non-toxic to humans. ${ }^{19}$ The DSHS, taking a conservative approach, estimates $10 \%$ of the total arsenic in any fish is inorganic arsenic and derives estimates of inorganic arsenic concentration in each fish by multiplying the reported total arsenic concentration in the sample by a factor of 0.1.

## Mercury

Nearly all mercury in upper trophic level fish three years of age or older is methylmercury. ${ }^{20}$ Thus, the total mercury concentration in a fish of legal size for possession in Texas serves well as a surrogate for methylmercury concentration. Because methylmercury analyses are difficult to perform accurately and are more expensive than total mercury analyses, the USEPA recommends that states determine total mercury concentration in a fish and that - to protect human health - states conservatively assume all reported mercury in fish or shellfish is methylmercury. The GERG laboratory thus analyzed fish tissues for total mercury. In its risk characterizations, the DSHS compares mercury concentrations in tissues to a comparison value derived from the Agency for Toxic Substances and Disease Registry's (ATSDR) minimal risk level (MRL) for methylmercury. ${ }^{21}$ (In these risk characterizations, the DSHS interchangeably utilizes the terms "mercury," "methylmercury," or "organic mercury" to refer to methylmercury in fish).

## Percent Lipids

The percent lipids content (wet weight basis) of a tissue sample is defined as the percent of material extracted from biological tissue with methylene chloride. ${ }^{22}$ Tissue samples were extracted with methylene chloride in the presence of sodium sulfate and an aliquot of the extract was removed for lipid determination, filtered and concentrated to a known volume. A subsample is removed, the solvent is evaporated, the lipid residue weighed, and the percent lipid content is determined.

## Polychlorinated Biphenyls (PCBs)

For PCBs, the USEPA suggests that each state measures congeners of PCBs in fish and shellfish rather than homologs ${ }^{\text {c }}$ or Aroclors ${ }^{\text {®d }}$ because the USEPA considers congener analysis the most sensitive technique for detecting PCBs in environmental media. ${ }^{23,}{ }^{20}$ Although only about 130 PCB congeners were routinely present in PCB mixtures manufactured and commonly used in the U.S. The GERG laboratory analyzes and reports the presence and concentrations of all 209 possible PCB congeners. From the congener analyses, the laboratory also computes and reports concentrations of PCB homologs and of Aroclor ${ }^{\circledR}$ mixtures. Despite the USEPA's suggestion that the states utilize PCB congeners rather than Aroclors ${ }^{\circledR}$ or homologs for toxicity estimates, the toxicity literature does not reflect state-of-the-art laboratory science. To accommodate this inconsistency, the DSHS utilizes recommendations from the National Oceanic and Atmospheric Administration (NOAA), ${ }^{24}$ from McFarland and Clarke, ${ }^{25}$ and from the USEPA's guidance documents for assessing contaminants in fish and shellfish. ${ }^{13,18}$ Based on evaluation of these recommendations, the DSHS selected 43 of 209 congeners to characterize "total" PCBs. The referenced authors chose to use congeners that were relatively abundant in the environment, were likely to occur in aquatic life, and likely to show toxic effects. SALG risk assessors summed the 43 congeners to derive "total" PCB concentration in each sample. SALG risk assessors then averaged the summed congeners within each group (e.g., fish species, sample site, or combination of species and site) to derive a mean PCB concentration for each group.

Using only a few PCB congeners to determine total PCB concentrations could underestimate PCB levels in fish tissue. Nonetheless, the method complies with expert recommendations on evaluation of PCBs in fish or shellfish. Therefore, SALG risk assessors compare average PCB concentrations of the 43 congeners with health assessment comparison (HAC) values derived from information on PCB mixtures held in the USEPA's Integrated Risk Information System (IRIS) database. ${ }^{26}$ IRIS currently contains noncarcinogenic toxicity information for three Aroclor ${ }^{\circ}$ mixtures: Aroclors ${ }^{\bullet} 1016,1248$, and 1254 . IRIS does not contain complete information for all mixtures. For instance, IRIS has derived reference doses (RfDs) for Aroclors 1016 and 1254. Aroclor 1016 was a commercial mixture produced in the latter years of commercial production of PCBs in the United States. Aroclor 1016 was a fraction of Aroclor 1254 that was supposedly devoid of dibenzofurans, in contrast to Aroclor $1254 .{ }^{27}$ Systemic toxicity estimates in the present document reflect comparisons derived from the USEPA's RfD for Aroclor 1254 because Aroclor 1254 contains many of the 43 congeners selected by McFarland and Clark and NOAA. As of yet, IRIS does not contain information on the systemic toxicity of individual PCB congeners.

[^2]For assessment of cancer risk from exposure to PCBs, the SALG uses the USEPA's highest slope factor of 2.0 milligram per kilogram per day ( $\mathrm{mg} / \mathrm{kg} /$ day) to calculate the probability of lifetime excess cancer risk from PCB ingestion. The SALG based its decision to use the most conservative slope factor available for PCBs on factors, such as food chain exposure; the presence of dioxinlike tumor-promoting or persistent congeners; and, the likelihood of early-life exposure. ${ }^{26}$

## Calculation of Dioxin Toxicity Equivalence (TEQ)

PCDDs/PCDFs are families of aromatic chemicals containing one to eight chlorine atoms. The molecular structures differ not only with respect to the number of chlorines on the molecule, but also with the positions of those chlorines on the carbon atoms of the molecule. The number and positions of the chlorines on the dibenzofuran or dibenzo-p-dioxin nucleus directly affects the toxicity of the various congeners. Toxicity increases as the number of chlorines increases to four chlorines, then decreases with increasing numbers of chlorine atoms - up to a maximum of eight. With respect to the position of chlorines on the dibenzo-p-dioxin/dibenzofuran nucleus, it appears that those congeners with chlorine substitutions in the $2,3,7$, and 8 positions are more toxic than congeners with chlorine substitutions in other positions. To illustrate, the most toxic form of PCDDs is 2,3,7,8-tetrachlorodibenzo-p-dioxin (2,3,7,8-TCDD), a 4-chlorine molecule having one chlorine substituted for hydrogen at each of the $2,3,7$, and 8 carbon positions on the dibenzo- $p$-dioxin. To gain some measure of toxic equivalence, $2,3,7,8$-TCDD assigned a toxicity equivalency factor (TEF) of 1.0 - is the standard against which other congeners are measured. Other congeners are given weighting factors, or TEFs, of 1.0 or less based on experiments comparing the toxicity of the congener relative to that of 2,3,7,8TCDD. ${ }^{28,29}$
Using this technique, the DSHS converted PCDD or PCDF congeners in each tissue sample from the present survey to toxic equivalent concentrations (TEQs) by multiplying each congener's concentration by its TEF, producing a dose roughly equivalent in toxicity to that of the same dose of $2,3,7,8$-TCDD. The total TEQ for any sample is the sum of the TEQs for each of the congeners in the sample, calculated according to the following formula. ${ }^{30}$

```
    n
Total TEQs = \Sigma(Cl x TEF)
    i=1
Cl = concentration of a given congener
TEF = toxicity equivalence factor for the given congener
n = # of congeners
i = initial congener
\Sigma= sum
```


## Derivation and Application of Health-Based Assessment Comparison Values for Systemic (Noncarcinogenic) Effects (HAC ${ }_{\text {nonca }}$ ) of Consumed Chemical Contaminants

The effects of exposure to any hazardous substance depend, among other factors, on the dose, the route of exposure, the duration of exposure, the manner in which the exposure occurs, the genetic makeup, personal traits and habits of the exposed, or the presence of other chemicals. ${ }^{31}$ People who regularly consume contaminated fish or shellfish conceivably suffer repeated low-dose exposures to contaminants in fish or shellfish over extended periods (episodic exposures to low doses). Such exposures are unlikely to result in acute toxicity but may increase risk of subtle, chronic, and/or delayed adverse health effects that may include: cancer, benign tumors; birth defects; infertility; blood disorders; brain damage; peripheral nerve damage; lung disease; and kidney disease. ${ }^{31}$

If diverse species of fish or shellfish are available, the SALG presumes that people eat a variety of species from a water body. Further, SALG risk assessors assume that most fish species are mobile. SALG risk assessors may combine data from different fish species and/or sample sites within a water body to evaluate mean contaminant concentrations of toxicants in all samples as a whole. This approach intuitively reflects consumers' likely exposure over time to contaminants in fish or shellfish from any water body but may not reflect the reality of exposure at a specific location within a water body or a single point in time. The DSHS reserves the right to project risks associated with ingestion of individual species of fish or shellfish from separate collection sites within a water body or at higher than average concentrations (e.g., the upper 95 percent confidence limit on the mean). The SALG evaluates contaminants in fish or shellfish by comparing the mean or the $95 \%$ upper confidence limit on the mean concentration of a contaminant to its HAC value (e.g., in $\mathrm{mg} / \mathrm{kg}$ ) for non-cancer or cancer endpoints. The mean is the preferred comparison statistic. However, the $95 \%$ upper confidence limit may be used when evaluating small sample sizes.

In deriving HAC values for systemic (noncarcinogenic; $\mathrm{HAC}_{\text {nonca }}$ ) effects, the SALG assumes a standard adult weighs 70 kilograms (kg) and consumes 30 g of fish or shellfish per day (about one eight-ounce meal per week) and uses the USEPA's RfD ${ }^{32}$ or the ATSDR's chronic oral MRLs. ${ }^{33}$ When RfDs or MRLs are not available the SALG may use a Food and Nutrition Board, Institute of Medicine, National Academies tolerable upper intake level (UL) for nutrients. ${ }^{e}$ The USEPA defines an RfD as

An estimate of a daily oral exposure for a given duration to the human population (including susceptible subgroups) that is likely to be without an appreciable risk of adverse health effects over a lifetime. ${ }^{34}$

The USEPA also states that the RfD

[^3]> ... is derived from a BMDL (benchmark dose lower confidence limit), a NOAEL (no observed adverse effect level), a LOAEL (lowest observed adverse effect level), or another suitable point of departure, with uncertainty/variability factors applied to reflect limitations of the data used. [Durations include acute, short-term, subchronic, and chronic and are defined individually in this glossary] and RfDs are generally reserved for health effects thought to have a threshold or a low dose limit for producing effects. ${ }^{34}$

The ATSDR uses a similar technique to derive its MRLs. ${ }^{33}$ The DSHS divides the estimated daily dose derived from the measured concentration in fish tissue by the contaminant's RfD or MRL to derive a hazard quotient (HQ). The USEPA defines an HQ as

> ...the ratio of the estimated exposure dose of a contaminant (mg/kg/day) to the contaminant's RfD or MRL ( $\mathrm{mg} / \mathrm{kg} /$ day). ${ }^{35}$

Note that, according to the USEPA, a linear increase in the HQ for a toxicant does not imply a linear increase in the likelihood or severity of systemic adverse effects. Thus, an HQ of 4.0 does not mean the concentration in the dose will be four times as toxic as that same substance would be if the HQ were equal to 1.0. An HQ of 4.0 also does not imply that adverse events will occur four times as often as if the HQ for the substance in question were 1.0. Rather, the USEPA suggests that an HQ or a hazard index (HI) - defined as the sum of HQs for contaminants to which an individual is exposed simultaneously - that computes to less than 1.0 should be interpreted as "no cause for concern" whereas, an HQ or HI greater than or equal to 1.0 "should indicate some cause for concern."

The SALG does not utilize HQs to determine the likelihood of occurrence of adverse systemic (noncarcinogenic) health effects. Instead, in a manner similar to the USEPA's decision process, the SALG may utilize computed HQs as a qualitative measurement. Qualitatively, HQs less than 1.0 are unlikely to be cause for concern while HQs greater than or equal to 1.0 might suggest the recommendation of a regulatory action to ensure protection of public health. Similarly, risk assessors at the DSHS may utilize an HQ to determine the need for further study of a water body's fauna. Notwithstanding the above discussion, the oral RfD derived by the USEPA represents chronic consumption. Thus, regularly eating fish containing a toxic chemical, the HQ of which is less than 1.0 is unlikely to cause adverse systemic health effects, whereas routine consumption of fish or shellfish in which the HQ equals or exceeds 1.0 represents a qualitatively unacceptable increase in the likelihood of systemic adverse health outcomes.

Although the DSHS utilizes chemical specific RfDs when possible, if an RfD is not available for a contaminant, the USEPA advises risk assessors to consider evaluating the contaminant by comparing it to the published RfD (or the MRL) of a contaminant of similar molecular structure or one with a similar mode or mechanism of action. For instance, Aroclor 1260 has no RfD, so the DSHS uses the reference dose for Aroclor 1254 to assess the likelihood of systemic (noncarcinogenic) effects of Aroclor 1260. ${ }^{33}$

In developing oral RfDs and MRLs, federal scientists review the extant literature to devise NOAELs, LOAELs, or benchmark doses (BMDs) from experimental studies. Uncertainty factors are then utilized to minimize potential systemic adverse health effects in people who are exposed through consumption of contaminated materials by accounting for certain conditions that may be undetermined by the experimental data. These include extrapolation from animals to humans (interspecies variability), intra-human variability, and use of a subchronic study rather than a chronic study to determine the NOAEL, LOAEL, or BMD, and database insufficiencies. ${ }^{32,34}$ Vulnerable groups such as women who are pregnant or lactating, women who may become pregnant, infants, children, people with chronic illnesses, those with compromised immune systems, the elderly, or those who consume exceptionally large servings are considered sensitive populations by risk assessors and USEPA. These sensitive groups also receive special consideration in calculation of an RfD. ${ }^{34}$

The primary method for assessing the toxicity of component-based mixtures of chemicals in environmental media is the HI . The USEPA recommends HI methodology for groups of toxicologically similar chemicals or chemicals that affect the same target organ. The HI for the toxic effects of a chemical mixture on a single target organ is actually a simulated HQ calculated as if the mixture were a single chemical. The default procedure for calculating the HI for the exposure mixture is to add the hazard quotients (the ratio of the external exposure dose to the RfD) for all the mixture's component chemicals that affect the same target organ (e.g., the liver). The toxicity of a particular mixture on the liver represented by the HI should approximate the toxicity that would have occurred were the observed effects caused by a higher dose of a single toxicant (additive effects). The components to be included in the HI calculation are any chemical components of the mixture that show the effect described by the HI , regardless of the critical effect from which the RfD came. Assessors should calculate a separate HI for each toxic effect.

Because the RfD is derived for the critical effect (the "toxic effect occurring at the lowest dose of a chemical"), an HI computed from HQs based on the RfDs for the separate chemicals may be overly conservative. That is, using RfDs to calculate HIs may overestimate health risks from consumption of specific mixtures for which no experimentally derived information is available. The USEPA states that
the HI is a quantitative decision aid that requires toxicity values as well as exposure estimates. When each organ-specific HI for a mixture is less than one and all relevant effects have been considered in the assessment, the exposure being assessed for potential systemic toxicity should be interpreted as unlikely to result in significant toxicity.

And

When any effect-specific HI exceeds one, concern exists over potential toxicity. As more HIs for different effects exceed one, the potential for human toxicity also increases.

Thus,

Concern should increase as the number of effect-specific HI's exceeding one increases. As a larger number of effect-specific HIs exceed one, concern over potential toxicity should also increase. As with HQs, this potential for risk is not the same as probabilistic risk; a doubling of the HI does not necessarily indicate a doubling of toxic risk.

## Derivation and Application of Health-Based Assessment Comparison Values for Application to the Carcinogenic Effects ( $\mathrm{HAC}_{c a}$ ) of Consumed Chemical Contaminants

The DSHS calculates cancer-risk comparison values ( $\mathrm{HAC}_{\mathrm{ca}}$ ) from the USEPA's chemical-specific cancer potency factors (CPFs), also known as cancer slope factors (CSFs), derived through mathematical modeling from carcinogenicity studies. For carcinogenic outcomes, the DSHS calculates a theoretical lifetime excess risk of cancer for specific exposure scenarios for carcinogens, using a standard $70-\mathrm{kg}$ body weight and assuming an adult consumes 30 grams of edible tissue per day. The SALG risk assessors incorporate two additional factors into determinations of theoretical lifetime excess cancer risk: (1) an acceptable lifetime risk level (ARL) ${ }^{34}$ of one excess cancer case in 10,000 persons whose average daily exposure is equivalent; and, (2) daily exposure for 30 years, a modification of the 70 -year lifetime exposure assumed by the USEPA. Comparison values used to assess the probability of cancer do not contain "uncertainty" factors. However, conclusions drawn from probability determinations infer substantial safety margins for all people by virtue of the models utilized to derive the slope factors (cancer potency factors) used in calculating the $\mathrm{HAC}_{c a}$.

Because the calculated comparison values (HAC values) are conservative, exceeding a HAC value does not necessarily mean adverse health effects will occur. The perceived strict demarcation between acceptable and unacceptable exposures or risks is primarily a tool used by risk managers along with other information to make decisions about the degree of risk incurred by those who consume contaminated fish or shellfish. Moreover, comparison values for adverse health effects do not represent sharp dividing lines (obvious demarcations) between safe and unsafe exposures. For example, the DSHS considers it unacceptable when consumption of four or fewer meals per month of contaminated fish or shellfish would result in exposure to contaminant(s) in excess of a HAC value or other measure of risk. The DSHS also advises people who wish to minimize exposure to contaminants in fish or shellfish to eat a variety of fish and/or shellfish and to limit consumption of those species most likely to contain toxic contaminants. The DSHS aims to protect vulnerable subpopulations with its consumption advice, assuming that advice protective of vulnerable subgroups will also protect the general population from potential adverse health effects associated with consumption of contaminated fish or shellfish.

## Children's Health Considerations

The DSHS recognizes that fetuses, infants, and children may be uniquely susceptible to the effects of toxic chemicals and suggests that exceptional susceptibilities demand special attention. ${ }^{36,37}$ Windows of special vulnerability (known as "critical developmental periods") exist during development. Critical periods occur particularly during early gestation (weeks 0 through 8) but can occur at any time during development (pregnancy, infancy, childhood, or adolescence) at times when toxicants can impair or alter the structure or function of susceptible systems. ${ }^{38}$ Unique early sensitivities may exist after birth because organs and body systems are structurally or functionally immature at birth, continuing to develop throughout infancy, childhood, and adolescence. Developmental variables may influence the mechanisms or rates of absorption, metabolism, storage, or excretion of toxicants. Any of these factors could alter the concentration of biologically effective toxicant at the target organ(s) or could modulate target organ response to the toxicant. Children's exposures to toxicants may be more extensive than adults' exposures because children consume more food and liquids in proportion to their body weights than adults consume. Infants can ingest toxicants through breast milk, an exposure pathway that often goes unrecognized. Nonetheless, the advantages of breastfeeding outweigh the probability of significant exposure to infants through breast milk and women are encouraged to continue breastfeeding and to limit exposure of their infants by limiting intake of the contaminated foodstuff. Children may experience effects at a lower exposure dose than might adults because children's organs may be more sensitive to the effects of toxicants. Stated differently, children's systems could respond more extensively or with greater severity to a given dose than would an adult organ exposed to an equivalent dose of a toxicant. Children could be more prone to developing certain cancers from chemical exposures than are adults. ${ }^{39}$ In any case, if a chemical or a class of chemicals is observed to be, or is thought to be, more toxic to fetuses, infants, or children, the constants (e.g., RfD, MRL, or CPF) are usually modified further to assure the immature systems' potentially greater susceptibilities are not perturbed. ${ }^{32}$ Additionally, in accordance with the ATSDR's Child Health Initiative ${ }^{40}$ and the USEPA's National Agenda to Protect Children's Health from Environmental Threats, ${ }^{41}$ the DSHS further seeks to protect children from the possible negative effects of toxicants in fish by suggesting that this potentially sensitive subgroup consume smaller quantities of contaminated fish or shellfish than adults consume. Thus, the DSHS recommends that children weighing 35 kg or less and/or who are 11 years of age or younger limit exposure to contaminants in fish or shellfish by eating no more than four-ounces per meal of the contaminated species. The DSHS also recommends that consumers spread these meals over time. For instance, if the DSHS issues consumption advice that recommends consumption of no more than two meals per month of a contaminated species, those children should eat no more than 24 four ounce meals of the contaminated fish or shellfish per year and should not eat such fish or shellfish more than twice per month.

## Data Analysis and Statistical Methods

The SALG risk assessors imported Excel ${ }^{\circledR}$ files into Systat ${ }^{\circledR}$ statistical software, version 13.1 installed on IBM-compatible microcomputers (Dell, Inc), to generate descriptive statistics
(mean, 95\% confidence limits of the arithmetic mean, standard deviation, median, minimum, and maximum concentrations) for reported chemical contaminants. ${ }^{42}$ In computing descriptive statistics, SALG risk assessors utilized $1 / 2$ the reporting limit (RL) for analytes designated as not detected (ND) or estimated (J-values). ${ }^{f}$ The SALG risk assessors calculated PCDDs/PCDFs descriptive statistics using estimated concentrations (J-values) and assuming zero for PCDDs/PCDFs designated as ND.g The change in methodology for computing PCDDs/PCDFs descriptive statistics is due to the proximity of the reporting limits to the HAC value. Assuming $1 / 2$ the RL for PCDDs/PCDFs designated as ND or J-values would unnecessarily overestimate the concentration of PCDDs/PCDFs in each fish tissue sample. The SALG used the descriptive statistics from the above calculations to produce the present report. The SALG employed Microsoft Excel ${ }^{\bullet}$ spreadsheets to create figures, to compute $\mathrm{HAC}_{\text {nonca }}$ and $\mathrm{HAC}_{c a}$ values for contaminants, and to calculate HQs, HIs, cancer risk probabilities, and meal consumption limits for fish from Ellison Creek Reservoir. ${ }^{43}$ When lead concentrations in fish or shellfish are high, SALG risk assessors may utilize the USEPA's Interactive Environmental Uptake Bio-Kinetic (IEUBK) model to determine whether consumption of lead-contaminated fish could cause a child's blood lead (PbB) level to exceed the Centers for Disease Control and Prevention's (CDC) lead concentration of concern in children's blood ( $5 \mathrm{mcg} / \mathrm{dL}$ ). ${ }^{44,45}$

The SALG risk assessors also performed other types of statistical analyses to evaluate dioxin and PCB data. Statistical significance was determined at $p \leq 0.05$ for all statistical analyses. When appropriate and as needed to meet assumptions of the statistical tests, the SALG risk assessors $\log _{e}$-transformed the data to improve normality and best fit. The SALG risk assessors performed linear correlation ( $r$ ) to describe associations between contaminant concentrations and total length (TL), fish age, and percent lipid composition. For those associations that were positive and significant, the SALG risk assessors performed linear regression analyses $\left(r^{2}\right)$ to measure the strength and further describe the relationships.

## RESULTS

The GERG laboratory completed analyses and electronically transmitted the results of the Ellison Creek Reservoir samples collected April-May 2014 to the SALG in August 2014. The laboratory reported the analytical results for metals, pesticides, PCBs, PCDDs/PCDFs, SVOCs, and VOCs.

For reference, Table 1 contains a list of fish samples collected by sample site. Tables 2.1-2.9 present the results of metals analyses. Tables 3 and 4.1-4.3 contain summary results for

[^4]pesticides and PCBs, respectively. Table 5.1-5.3 summarizes the PCDD/PCDF analyses. Table 6 depicts summary results for VOCs (i.e., trichlorofluoromethane). This report does not display SVOC data because these contaminants were not present at concentrations of concern in fish collected from Ellison Creek Reservoir during the described survey. Unless otherwise stated, table summaries present the number of samples with detected concentrations of contaminants, the number of samples tested, the mean concentration and standard deviation, and the minimum and the maximum concentrations. In the tables, results may be reported as ND, below detection limit (BDL) for estimated concentrations or "J-values", or as concentrations at or above the reporting limit ( $R L$ ).

## Inorganic Contaminants

## Arsenic, Cadmium, Copper, Lead, Selenium, and Zinc

The GERG laboratory analyzed a subset of 100 fish tissue samples for six inorganic contaminants and 108 samples for mercury. All fish tissue samples from Ellison Creek Reservoir contained concentrations of arsenic, copper, mercury, selenium, and zinc (Tables 2.1-2.9).

The SALG evaluated three toxic metalloids having no known human physiological function (arsenic, cadmium, and lead) in the samples collected from Ellison Creek Reservoir. Arsenic concentrations ranged from BDL to $0.445 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.074 \pm 0.067 \mathrm{mg} / \mathrm{kg}$ (Table 2.1). Ninety-three of 100 fish analyzed contained estimated concentrations below the RL for cadmium (Table 2.2). Lead concentrations ranged from ND to $1.011 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.258 \pm 0.154 \mathrm{mg} / \mathrm{kg}$ (Table 2.4).

Three of the metalloids analyzed are essential trace elements: copper, selenium, and zinc. Copper concentrations ranged from BDL to $2.714 \mathrm{mg} / \mathrm{kg}$ (Table 2.3). All fish tissue samples contained selenium. Selenium concentrations ranged from BDL to $4.477 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.429 \pm 0.797 \mathrm{mg} / \mathrm{kg}$ (Table 2.5). All samples also contained zinc. The mean zinc concentration in fish tissue samples from Ellison Creek Reservoir was $4.122 \pm 1.531 \mathrm{mg} / \mathrm{kg}$ (Table 2.6).

## Mercury

All fish tissue samples evaluated from Ellison Creek Reservoir contained mercury (Tables 2.7-2.9). Mercury concentrations ranged from 0.009 to $0.527 \mathrm{mg} / \mathrm{kg}$. The mean mercury concentration for the 108 fish tissue samples analyzed was $0.078 \pm 0.084 \mathrm{mg} / \mathrm{kg}$.

## Organic Contaminants

## Pesticides

All samples examined contained concentrations of chlordane and 4,4'dichlorodiphenyldichloroethylene (DDE). Chlordane concentrations ranged from 0.0010 to $0.0216 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.0068 \pm 0.0055 \mathrm{mg} / \mathrm{kg}$ (Table 3). DDE ranged from 0.0007 to
$0.0117 \mathrm{mg} / \mathrm{kg}$ with a mean $0.0041 \pm 0.0029 \mathrm{mg} / \mathrm{kg}$ (Table 3). The mean pentachloroanisole concentration in fish tissue samples from Ellison Creek Reservoir was $0.0007 \pm 0.0009$ (Table 3). Estimated to low concentrations greater than the reporting limit of hexachlorobenzene, heptachlor epoxide, and methoxychlor were present in one or more fish samples (data not presented). Estimated concentrations were reported for 4,4'- dichlorodiphenyldichloroethane (DDD), dacthal, mirex, pentachlorobenzene, and tetrachlorobenzene (data not presented).

## PCBS

All fish tissue samples evaluated from Ellison Creek Reservoir contained PCBs (Tables 4.1-4.3). One redbreast sunfish sample was removed from data analysis for failure to meet the contract required reporting limit for PCBs. Across all sample sites and species, PCB concentrations ranged from 0.009 (black crappie) to $0.423 \mathrm{mg} / \mathrm{kg}$ (common carp). The mean PCB concentration for the 107 fish tissue samples evaluated was $0.102 \pm 0.085 \mathrm{mg} / \mathrm{kg}$. PCB concentrations in fish appeared to be positively related to TL and percent lipids ( $r^{2}=0.263, n=107, p<0.0005$; $r^{2}=$ $0.470, n=107, p<0.0005$; Figures $2-3$ ).

## Bowfin

Three bowfin ranging from 20.7 to 27.6 inches TL ( $\bar{X}-24.0$ inches TL ) were analyzed for PCBs (Table 1). Currently, there is no minimum length limit for bowfin in Texas waters. ${ }^{46}$ PCBs concentrations ranged from 0.010 to $0.048 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.027 \pm 0.019 \mathrm{mg} / \mathrm{kg}$ (Tables 4.1-4.3).

## Channel catfish

Thirty-five channel catfish ranging from 14.8 to 24.8 inches $\mathrm{TL}(\bar{X}-19.2$ inches TL ) and from three to 13 years of age were analyzed for PCBs (Table 1; Figure 4). One-hundred percent of the channel catfish samples examined were of legal size ( $\geq 12$ inches TL ). ${ }^{46}$ PCB concentrations ranged from 0.027 to $0.408 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.111 \pm 0.084 \mathrm{mg} / \mathrm{kg}$ (Tables $4.1-4.3$ ). PCB concentrations in channel catfish appeared to be positively related to TL and age ( $r^{2}=0.483, n=$ $35, p<0.0005 ; r^{2}=0.289, n=35, p=0.0009$; Figures 5-6). There was no apparent correlation between PCB concentration and percent lipids ( $r=0.241, n=35, p=0.163$ ).

## Common carp

Thirteen common carp ranging from 23.1 to 29.1 inches TL ( $\bar{X}-26.0$ inches TL ) were analyzed for PCBs (Table 1). Currently, there is no minimum length limit for common carp in Texas waters. ${ }^{46}$ PCB concentrations ranged from 0.098 to $0.423 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.209 \pm 0.086$ $\mathrm{mg} / \mathrm{kg}$ (Tables 4.1-4.3).

## Crappie Spp. (black and white)

Seven crappie ranging from 10.2 to 13.7 inches TL ( $\bar{X}-11.8$ inches TL ) and from three to six years of age were analyzed for PCBs (Table 1; Figure 7). One-hundred percent of the crappie samples examined were of legal size ( $\geq 10$ inches TL ). ${ }^{46}$ PCB concentrations ranged from 0.009 to $0.048 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.022 \pm 0.013 \mathrm{mg} / \mathrm{kg}$ (Tables $4.1-4.3$ ).

## Flathead catfish

Eight flathead catfish ranging from 18.3 to 35.4 inches TL ( $\bar{X}-27.6$ inches TL ) and from four to 13 years of age were analyzed for PCBs (Table 1; Figure 8). One-hundred percent of the flathead catfish samples examined were of legal size ( $\geq 18$ inches TL ). ${ }^{46} \mathrm{PCB}$ concentrations ranged from 0.045 to $0.229 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.095 \pm 0.065 \mathrm{mg} / \mathrm{kg}$ (Tables 4.1-4.3).

## Hybrid striped bass

Two hybrid striped bass ranging from 24.4 to 24.5 inches $\mathrm{TL}(\bar{X}-24.5$ inches TL ) and nine years of age were analyzed for PCBs (Table 1). One-hundred percent of the hybrid striped bass samples examined were of legal size ( $\geq 18$ inches TL ). ${ }^{46} \mathrm{PCB}$ concentrations ranged from 0.190 to $0.357 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.273 \pm 0.118 \mathrm{mg} / \mathrm{kg}$ (Tables $4.1-4.3$ ).

## Largemouth bass

Thirty largemouth bass ranging from 14.7 to 22.8 inches TL ( $\bar{X}-18.3$ inches TL ) and from three to nine years of age were analyzed for PCBs (Table 1; Figure 9). One-hundred percent of the largemouth samples examined were of legal size ( $\geq 14$ inches TL ). ${ }^{46}$ PCB concentrations ranged from 0.018 to $0.227 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.064 \pm 0.038 \mathrm{mg} / \mathrm{kg}$ (Tables $4.1-4.3$ ). PCB concentrations in largemouth bass appeared to be positively related to percent lipids ( $r^{2}=$ $0.473, n=30, p<0.0005$; Figure 10). There was no apparent correlation between PCB concentration and TL and age ( $r=0.187, n=30, p=0.322 ; r=0.150, n=30, p=0.431$ ).

## Spotted gar

Four spotted gar ranging from 25.5 to 28.1 inches TL ( $\bar{X}-27.0$ inches TL ) were analyzed for PCBs (Table 1). Currently, there is no minimum length limit for spotted gar in Texas waters. ${ }^{46}$ PCB concentrations ranged from 0.036 to $0.179 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.111 \pm 0.061 \mathrm{mg} / \mathrm{kg}$ (Tables 4.1-4.3).

## Sunfish Spp. (bluegill, green, and redbreast)

Three sunfish ranging from 7.2 to 8.9 inches $\mathrm{TL}(\bar{X}-8.3$ inches TL ) and from three to seven years of age were analyzed for PCBs (Table 1). Currently, there is no minimum length limit for any sunfish species in Texas waters. ${ }^{46}$ PCB concentrations ranged from 0.012 to $0.162 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.103 \pm 0.080 \mathrm{mg} / \mathrm{kg}$ (Tables 4.1-4.3).

## White bass

Two white bass ranging from 24.4 to 24.5 inches TL ( $\bar{X}-24.5$ inches $T L$ ) and from four to five years of age were analyzed for PCBs (Table 1). One-hundred percent of the white bass samples examined were of legal size ( $\geq 10$ inches TL ). ${ }^{46}$ PCB concentrations ranged from 0.058 to 0.088 $\mathrm{mg} / \mathrm{kg}$ with a mean of $0.073 \pm 0.021 \mathrm{mg} / \mathrm{kg}$ (Tables $4.1-4.3$ ).

## PCDDs/PCDFs

Ninety-eight of 100 fish tissue samples contained at least one of the 17 PCDD/PCDF congeners ranging from ND-10.400 TEQ $\mathrm{pg} / \mathrm{g}$ with a mean of $1.013 \pm 1.340$ TEQ $\mathrm{pg} / \mathrm{g}$ and a median of 0.634 TEQ pg/g (Table 5.1-5.3). No samples contained all 17 congeners (data not shown). Hybrid striped bass contained the highest mean PCDD/PCDF TEQ concentration ( $3.803 \pm 3.005 \mathrm{pg} / \mathrm{g}$; Table 5.3). PCDD/PCDF TEQ concentrations in fish appeared to be positively related to TL and percent lipids ( $r^{2}=0.103, n=100, p=0.001 ; r^{2}=0.196, n=100, p<0.0005$; Figures 11-12).

## Bowfin

PCDD/PCDF TEQ concentrations ranged from 0.031 to $0.541 \mathrm{pg} / \mathrm{g}$ with a mean of $0.216 \pm 0.282$ $\mathrm{pg} / \mathrm{g}$ and median $0.076 \mathrm{pg} / \mathrm{g}(n=3$; Tables 5.1-5.3).

## Channel catfish

PCDD/PCDF TEQ concentrations ranged from 0.142 to $10.400 \mathrm{pg} / \mathrm{g}$ with a mean of $1.323 \pm 1.895$ $\mathrm{pg} / \mathrm{g}$ and median $0.676 \mathrm{pg} / \mathrm{g}(n=33$; Tables 5.1-5.3). PCDD/PCDF TEQ concentration in channel catfish appeared to be positively related to TL and percent lipid ( $r^{2}=0.410, n=33, p=0.0001 ; r^{2}$ $=0.135, n=33, p=0.04$; Figures $13-14)$. There was no apparent correlation between PCB concentration and age ( $r=0.265, n=33, p=0.137$ ).

## Common carp

PCDD/PCDF TEQ concentrations ranged from 0.575 to $3.241 \mathrm{pg} / \mathrm{g}$ with a mean of $1.575 \pm 0.917$ $\mathrm{pg} / \mathrm{g}$ and median $1.401 \mathrm{pg} / \mathrm{g}(n=13$; Tables 5.1-5.3).

## Crappie Spp. (black and white)

PCDD/PCDF TEQ concentrations ranged from ND to $0.294 \mathrm{pg} / \mathrm{g}$ with a mean of $0.139 \pm 0.121$ $\mathrm{pg} / \mathrm{g}$ and median $0.123 \mathrm{pg} / \mathrm{g}(n=7$; Tables 5.1-5.3).

## Flathead catfish

PCDD/PCDF TEQ concentrations ranged from 0.260 to $1.689 \mathrm{pg} / \mathrm{g}$ with a mean of $0.842 \pm 0.456$ $\mathrm{pg} / \mathrm{g}$ and median $0.716 \mathrm{pg} / \mathrm{g}(n=8$; Tables 5.1-5.3).

## Hybrid striped bass

PCDD/PCDF TEQ concentrations ranged from 1.678 to $5.928 \mathrm{pg} / \mathrm{g}$ with a mean of $3.803 \pm 3.005$ $\mathrm{pg} / \mathrm{g}$ and median $3.803 \mathrm{pg} / \mathrm{g}(n=2$; Tables 5.1-5.3).

## Largemouth bass

PCDD/PCDF TEQ concentrations ranged from ND to $1.927 \mathrm{pg} / \mathrm{g}$ with a mean of $0.556 \pm 0.432$ $\mathrm{pg} / \mathrm{g}$ and median $0.427 \mathrm{pg} / \mathrm{g}(n=28$; Tables 5.1-5.3). There was no apparent correlation between PCDD/PCDF TEQ concentration and TL, age, and percent lipids, respectively ( $r=0.361$, $n=28, p=0.06 ; r=-0.010, n=28, p=0.960 ; r=0.337, n=28, p=0.079$ ).

## Spotted gar

PCDD/PCDF TEQ concentrations ranged from 0.253 to $1.416 \mathrm{pg} / \mathrm{g}$ with a mean of $0.969 \pm 0.507$ $\mathrm{pg} / \mathrm{g}$ and median $1.104 \mathrm{pg} / \mathrm{g}(n=4$; Tables 5.1-5.3).

## White bass

PCDD/PCDF TEQ concentrations ranged from 0.769 to $1.045 \mathrm{pg} / \mathrm{g}$ with a mean of $0.907 \pm 0.195$ $\mathrm{pg} / \mathrm{g}$ and median $0.907 \mathrm{pg} / \mathrm{g}(n=2$; Tables 5.1-5.3).

## SVOCs

The GERG laboratory analyzed a subset of 16 Ellison Creek Reservoir fish tissue samples for SVOCs. Quantifiable concentrations greater than the reporting limit were reported for 7-12Dimethylbenz(a)anthracene in one fish sample (data not presented). Estimated concentrations of benzo(g,h,i)perylene, dibenz(a,j)acridine, dibenz(a,h)anthracene, diethyl phthalate, di-n-butyl phthalate, bis (2-ethylhexyl) phthalate, di-n-octyl phthalate, acetophenone, ethyl methaneasulfonate, and 2,6-dinitrotoluene were present in one or more fish samples analyzed (data not presented). The laboratory detected no other SVOCs in fish from Ellison Creek Reservoir.

## VOCs

The Seafood and Aquatic Life Group Survey Team Standard Operating Procedures and Quality Control/Assurance Manual contain a complete list of the 70 VOCs selected for analysis. A subset of 16 fish tissue samples were selected for analysis from Ellison Creek Reservoir. Trichlorofluoromethane concentrations ranged from ND $-0.332 \mathrm{mg} / \mathrm{kg}$ with a mean of $0.061 \pm 0.080 \mathrm{mg} / \mathrm{kg}$ (Table 6). Quantifiable concentrations greater than the reporting limit were reported for acetone, methylene chloride, and naphthalene in one or more fish samples (data not presented in tables). Estimated quantities of 2-butanone, m+p-xylene, o-xylene, and toluene were also present in one or more fish tissue samples analyzed from Ellison Creek Reservoir (data not presented).

Acetone, methylene chloride, and naphthalene were also identified in one or more of the procedural blanks, suggesting that that these compounds were introduced during sample preparation. VOC concentrations less than the reporting limit are difficult to interpret due to their uncertainty and may represent a false positive. The presence of many VOCs at concentrations less than the reporting limit may be the result of incomplete removal of the calibration standard from the adsorbent trap, so they are observed in the blank. VOC analytical methodology requires that the VOCs be thermally released from the adsorbent trap, transferred to the gas chromatograph (GC), and into the mass spectrometer (MS) for quantification.

## DISCUSSION

## Risk Characterization

Because variability and uncertainty are inherent to quantitative assessment of risk, the calculated risks of adverse health outcomes from exposure to toxicants can be orders of magnitude above or below actual risks. Variability in calculated and in actual risk may depend upon factors such as the use of animal instead of human studies, use of subchronic rather than chronic studies, interspecies variability, intra-species variability, and database insufficiency. Because most factors used to calculate comparison values result from experimental studies conducted in the laboratory on nonhuman subjects, variability and uncertainty might arise from the study chosen as the "critical" one, the species/strain of animal used in the critical study, the target organ selected as the "critical organ," exposure periods, exposure route, doses, or uncontrolled variations in other conditions. ${ }^{32}$ Despite such limitations, risk assessors must calculate parameters to represent potential toxicity to humans who consume contaminants in fish and other environmental media. The DSHS calculated risk parameters for noncarcinogenic and carcinogenic endpoints in those who would consume fish from the Ellison Creek Reservoir. Conclusions and recommendations predicated upon the stated goal of the DSHS to protect human health follow the discussion of the relevance of findings to risk.

## Characterization of Systemic (Noncarcinogenic) Health Effects from Consumption of Fish from Ellison Creek Reservoir

## Inorganic Contaminants

None of the species of fish evaluated contained arsenic, cadmium, copper, lead, mercury, selenium, or zinc at concentrations that equaled or exceeded DSHS guidelines for protection of human health or would likely cause systemic (noncancerous) risk to human health from consumption of fish from Ellison Creek Reservoir.

## Organic Contaminants

PCBs and PCDDs/PCDFs were observed in fish from Ellison Creek Reservoir at concentrations at or above their respective $\mathrm{HAC}_{\text {nonca }}(0.047 \mathrm{mg} / \mathrm{kg} ; 2.330 \mathrm{pg} / \mathrm{g}$; Tables $4.1-4.3,5.1-5.3$, and 8.18.3). None of the species of fish evaluated contained any other organic contaminants at concentrations that equaled or exceeded DSHS guidelines for protection of human health or would likely cause systemic (noncancerous) risk to human health from consumption of fish from Ellison Creek Reservoir.

## PCBs

All fish tissue samples ( $n=107$ ) evaluated contained PCBs. Seventy-seven percent of all samples analyzed contained PCB concentrations equaling or exceeding the HAC nonca for PCBs ( 0.047 $\mathrm{mg} / \mathrm{kg}$; Tables 4.1-4.3). Nine (bluegill, channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, redbreast sunfish, spotted gar, and white bass) of the 13 species of fish evaluated had mean PCB concentrations exceeding the HAC nonca for PCBs or an HQ of 1.0 (Tables 4.1-4.3 and 8.1-8.3). The all fish combined mean PCB concentration ( $0.102 \mathrm{mg} / \mathrm{kg}$ ) exceeded the HAC nonca for PCBs or an HQ of 1.0. PCB concentrations were positively related to TL and percent lipids indicating that PCB concentrations increase as fish grow and as their body fat increases (Figures 2-3). People should consider these relationships when choosing the size and species of fish they consume. The consumption of bluegill, channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, redbreast sunfish, spotted gar, and white bass from Ellison Creek Reservoir may pose potential systemic (noncancerous) health risks.

Meal consumption calculations are useful for risk managers to make fish consumption recommendations and/or take regulatory action. The SALG risk assessors calculated the number of eight-ounce meals of fish from Ellison Creek Reservoir that healthy adults could consume without significant risk of PCB-related adverse systemic effects (Tables 8.1-8.3). Meal consumption rates were based on the overall mean PCB concentration by species. The SALG risk assessors estimated that healthy adults could consume less than one eight-ounce meal per week for these species of fish: 0.4 meals per week of channel catfish; or, 0.2 meals per week of common carp; 0.5 meals per week of flathead catfish; 0.2 meals per week of hybrid striped bass; 0.7 meals per week of largemouth bass; 0.4 meals per week of spotted gar; 0.4 meals per week of sunfish Spp.; or, 0.6 meals per week of white bass. The SALG risk assessors suggest that
fish from Ellison Creek Reservoir contain PCBs at concentrations that may pose potential systemic (noncancerous) health risks and that people should not consume common carp and hybrid striped bass and limit their consumption of channel catfish, flathead catfish, largemouth bass, spotted gar, sunfish Spp., and white bass from Ellison Creek Reservoir. Because the developing nervous system of the human fetus and young children may be especially susceptible to adverse systemic (noncancerous) health effects associated with consuming PCBcontaminated fish, the SALG risk assessors recommend more conservative consumption guidance for this sensitive subpopulation.

## PCDDs/PCDFs

Ninety-eight of 100 fish tissue samples assayed contained PCDDs/PCDFs. Nine percent of all samples analyzed contained PCDD/PCDF concentrations exceeding the HAC nonca for PCDDs/PCDFs ( $2.330 \mathrm{pg} / \mathrm{g}$; Tables 5.1-5.3 and 8.1-8.3). One (hybrid striped bass) of 10 species of fish evaluated had mean PCDD/PCDF concentrations exceeding the $H A C_{\text {nonca }}$ for PCDDs/PCDFs or an HQ of 1.0 (Tables 5.1-5.3 and 8.1-8.3). The all fish combined mean PCDD/PCDF concentration did not exceed the HAC nonca for PCDDs/PCDFs or an HQ of 1.0. The consumption of hybrid striped bass from Ellison Creek Reservoir may pose potential systemic (noncancerous) health risks.

Meal consumption calculations are useful for risk managers to make fish consumption recommendations and/or take regulatory action. The SALG risk assessors calculated the number of eight-ounce meals of fish from Ellison Creek Reservoir that healthy adults could consume without significant risk of PCDD/PCDF -related adverse systemic effects (Tables 8.18.3). Meal consumption rates were based on the overall mean PCDD/PCDF concentration by species. The SALG risk assessors estimated that healthy adults could consume less than one eight-ounce meal per week of hybrid striped bass. The SALG risk assessors estimated that people should not consume more than 0.6 meals per week. The SALG risk assessors suggest that hybrid striped bass from Ellison Creek Reservoir contain PCDDs/PCDFs at concentrations that may pose potential systemic (noncancerous) health risks and that people should limit their consumption of hybrid striped bass from Ellison Creek Reservoir. Because the developing nervous system of the human fetus and young children may be especially susceptible to adverse systemic health effects associated with consuming PCDD/PCDF-contaminated fish, the SALG risk assessors recommend more conservative consumption guidance for this sensitive subpopulation.

## Characterization of Theoretical Lifetime Excess Cancer Risk from Consumption of Fish from Ellison Creek Reservoir

The USEPA classifies arsenic, most chlorinated pesticides, PCBs, and PCDDs/PCDFs as human carcinogens. Arsenic, chlordane, and DDE were present in fish samples analyzed from Ellison Creek Reservoir, but none of these contaminants evaluated singly by species or all species combined had mean contaminant concentrations that would be likely to increase the risk of
cancer to exceed the DSHS guideline for protection of human health of one excess cancer in 10,000 equally exposed individuals.

## PCBs

The mean PCB concentrations observed in hybrid striped bass exceed the DSHS guideline for protection of human health of one excess cancer in 10,000 equally exposed individuals and the $\mathrm{HAC}_{\mathrm{ca}}$ for PCBs ( $0.272 \mathrm{mg} / \mathrm{kg}$; Tables 4.1-4.3 and 9.1-9.4). PCB concentrations that equaled or exceeded the $\mathrm{HAC}_{\mathrm{ca}}$ for PCBs were observed in one or more samples of channel catfish and common carp. The all fish combined mean PCB concentration did not exceed the HAC $\mathrm{ca}_{\mathrm{a}}$ for PCBs.

The SALG risk assessors calculated the number of eight-ounce meals of hybrid striped bass from Ellison Creek Reservoir that healthy adults could consume without significantly increasing their lifetime excess cancer risk (Table 9.1-9.4). The SALG risk assessors estimated that healthy adults could consume less than one eight-ounce meal per week of hybrid striped bass ( 0.9 meals per week). Because children may experience effects at a lower exposure dose than adults, the SALG risk assessors recommend more conservative consumption guidance for this sensitive subpopulation. The SALG risk assessors suggest that consumption of hybrid striped bass from Ellison Creek Reservoir would be likely to increase the risk of cancer to exceed the DSHS guideline for protection of human health from PCB exposure.

## PCDDs/PCDFs

The mean PCDD/PCDF concentrations observed in hybrid striped bass exceed the DSHS guideline for protection of human health of one excess cancer in 10,000 equally exposed individuals or the $\mathrm{HAC}_{\mathrm{ca}}$ for PCDDs/PCDFs ( $3.490 \mathrm{pg} / \mathrm{g}$; Tables 5.1-5.3 and 9.1-9.4). The all fish combined mean PCDD/PCDF concentration did not exceed the HAC ${ }_{c a}$ for PCDDs/PCDFs. The consumption of hybrid striped bass from Ellison Creek Reservoir would be likely to increase the risk of cancer to exceed the DSHS guideline for protection of human health.

The SALG risk assessors calculated the number of eight-ounce meals of hybrid striped bass from Ellison Creek Reservoir that healthy adults could consume without significantly increasing their lifetime excess cancer risk (Tables 9.1-9.4). The SALG risk assessors estimated that healthy adults could consume less than one eight-ounce meal per week of hybrid striped bass ( 0.8 meals per week). Because children may experience effects at a lower exposure dose than might adults because children's systems may be more sensitive to the effects of toxicants, the SALG risk assessors recommend more conservative consumption guidance for this sensitive subpopulation. The SALG risk assessors suggest that consumption of hybrid striped bass from Ellison Creek Reservoir would be likely to increase the risk of cancer to exceed the DSHS guideline for protection of human health from PCDD/PCDF exposure.

## Characterization of Calculated Cumulative Systemic (Noncarcinogenic) Health Effects and of Cumulative Excess Lifetime Cancer Risk from Consumption of Fish from Ellison Creek Reservoir

## Cumulative Systemic (Noncarcinogenic) Health Effects

Cumulative systemic (noncarcinogenic) effects of toxicants may occur if more than one contaminant acts upon the same target organ or acts by the same mode or mechanism of action. PCBs and PCDDs/PCDFs in fish from Ellison Creek Reservoir could have these properties, especially with respect to effects on the immune system. Multiple organic contaminants in Ellison Creek Reservoir fish increased the likelihood of systemic adverse health outcomes for all species of fish evaluated (Tables 8.1-8.3). The combined toxicity of PCBs and PCDDs/PCDFs in channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, Sunfish Spp., and white bass exceeded an HI of 1.0.

Meal consumption calculations are useful for risk managers to make fish consumption recommendations and/or take regulatory action. The SALG risk assessors calculated the number of eight-ounce meals of fish from Ellison Creek Reservoir that healthy adults could consume without significant risk of PCB and/or PCDD/PCDF -related adverse systemic effects (Tables 8.1-8.3). Meal consumption rates were based on cumulative toxicity from exposure to PCBs and PCDDs/PCDFs by species. The SALG risk assessors estimated that healthy adults could consume less than one eight-ounce meal per week of channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, Sunfish Spp., or white bass (Tables 8.1-8.3). The SALG risk assessors suggest that channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, Sunfish Spp., and white bass from Ellison Creek Reservoir contain PCBs and PCDDs/PCDFs at concentrations that may pose potential systemic (noncarcinogenic) health risks and that people should limit their consumption of fish from Ellison Creek Reservoir. Because the developing body systems of the human fetus and young children may be especially susceptible, the SALG risk assessors recommend more conservative consumption guidance for these sensitive subpopulations.

## Cumulative Carcinogenic Health Effects

The SALG also queried the probability of increasing lifetime excess cancer risk from consuming fish containing multiple inorganic and organic contaminants. In most assessments of cancer risk from environmental exposures to chemical mixtures, researchers have considered any increase in cancerous or benign growths in one or more organs as cumulative, no matter the mode or mechanism of action of the contaminant. In this assessment, risk assessors added the calculated carcinogenic effect of arsenic, chlorinated pesticides, PCBs, and PCDFs/PCDDs (Tables 9.1-9.4). In each instance, addition of the cancer risk for these chemicals increased the theoretical lifetime excess cancer risk. The cancer risk increase did elevate lifetime excess cancer risk to a level greater than the DSHS guideline for protection of human health of one excess cancer in 10,000 persons equivalently exposed for channel catfish, common carp, and hybrid striped bass.

The consumption of channel catfish, common carp, and hybrid striped bass from Ellison Creek Reservoir likely increases the risk of cancer to exceed the DSHS guideline for protection of human health. The SALG risk assessors estimated that healthy adults could consume less than one eight-ounce meal per week for these species of fish: 0.9 meals per week of channel catfish; 0.7 meals per week of or common carp; or, 0.4 meals per week of hybrid striped bass (Tables 9.1-9.4). Because children may experience effects at a lower exposure dose than adults, the SALG risk assessors recommend more conservative consumption guidance for this sensitive subpopulation. The SALG risk assessors suggest that consumption of channel catfish, common carp, and hybrid striped bass from Ellison Creek Reservoir would be likely to increase the risk of cancer to exceed the DSHS guideline for protection of human health from multiple contaminant exposures.

## CONCLUSIONS

The SALG risk assessors prepare risk characterizations to determine public health hazards from consumption of fish and shellfish harvested from Texas water bodies by recreational or subsistence fishers. If necessary, the SALG risk assessors may suggest strategies for reducing risk to the health of those who may eat contaminated fish or seafood to risk managers at the DSHS, including the Texas Commissioner of Health.

This study addressed the public health implications of consuming fish from Ellison Creek Reservoir, located in Morris County, Texas. Risk assessors from the SALG conclude from the present characterization of potential adverse health effects from consuming fish from Ellison Creek Reservoir that:

1. Confidence in the conclusions for many species of fish is limited by the small sample size. Sampling a small number of fish (i.e., individual species of fish or all fish species combined) decreases the confidence of mean contaminant concentrations for the fish population thus adding uncertainty to the conclusions.
2. Black crappie, bowfin, channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, white bass, and white crappie mean concentrations of arsenic, cadmium, copper, lead, mercury, selenium, zinc, most pesticides, SVOCs, or VOCs; either singly, or in combination do not exceed the DSHS guidelines for protection of human health. Therefore, consumption of these species of fish from Ellison Creek Reservoir containing the above-listed contaminants poses no apparent risk to human health.
3. Black crappie, bowfin, green sunfish, and white crappie mean PCB concentrations do not exceed the DSHS guidelines for protection of human health for women past childbearing age and males 12 and older. Therefore, consumption of these species of fish from Ellison Creek Reservoir containing only PCBs poses no apparent risk to human health.
4. Black crappie, bowfin, channel catfish, common carp, flathead catfish, largemouth bass, spotted gar, white bass, and white crappie mean PCDD/PCDF TEQ concentrations do not exceed the DSHS guidelines for protection of human health for women past childbearing age and males 12 and older. Therefore, consumption of these species of fish from Ellison Creek Reservoir containing only PCDDs/PCDFs poses no apparent risk to human health.
5. Bluegill, channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, redbreast sunfish, spotted gar, and white bass mean PCB concentrations exceed the DSHS guidelines for protection of human health for women past childbearing age and males 12 and older. Regular or long-term consumption of these species of fish may result in adverse systemic (noncarcinogenic) health effects and/or increase the likelihood of carcinogenic health risks. Therefore, consumption of these species of fish from Ellison Creek Reservoir poses an apparent risk to human health.
6. Hybrid striped bass mean PCDD/PCDF TEQ concentrations exceed the DSHS guidelines for protection of human health for women past childbearing age and males 12 and older. Regular or long-term consumption of hybrid striped bass may result in adverse systemic (noncarcinogenic) health effects and/or increase the likelihood of carcinogenic health risks. Therefore, consumption of hybrid striped bass from Ellison Creek Reservoir poses an apparent risk to human health.
7. Consumption of multiple organic contaminants (i.e., PCDDs/PCDFs and PCBs) in channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, and white bass increases the likelihood of systemic (noncarcinogenic) health risks. Regular or long-term consumption of these species of fish may result in adverse systemic (noncarcinogenic) health effects. Therefore, consumption of these species of fish from Ellison Creek Reservoir poses an apparent risk to human health.
8. Consumption of multiple inorganic and/or organic contaminants observed in channel catfish, and hybrid striped bass increases the likelihood of carcinogenic health risks. Therefore, consumption of these species of fish from Ellison Creek Reservoir containing multiple contaminants poses an apparent risk to human health.

## RECOMMENDATIONS

Risk managers at the DSHS have established criteria for issuing fish consumption advisories based on approaches suggested by the USEPA. ${ }^{13,18,47}$ Risk managers at the DSHS may decide to take action to protect public health if a risk characterization confirms that people can eat four or fewer meals per month (adults: eight-ounces per meal; children: four-ounces per meal) of fish or shellfish from a water body under investigation. Risk management recommendations may be in the form of consumption advice or a ban on possession of fish from the affected water body. Fish or shellfish possession bans are enforceable under subchapter D of the Texas Health and Safety Code, part 436.061(a). ${ }^{48}$ Declarations of prohibited harvesting areas are
enforceable under the Texas Health and Safety Code, Subchapter D, parts 436.091 and 436.101. ${ }^{48}$ The DSHS consumption advice carries no penalty for noncompliance. Consumption advisories, instead, inform the public of potential health hazards associated with consuming contaminated fish or shellfish from Texas waters. With this information, people can make informed decisions about whether and/or how much, contaminated fish or shellfish, they wish to consume. The SALG concludes from this risk characterization that consuming channel catfish, common carp, flathead catfish, hybrid striped bass, largemouth bass, spotted gar, sunfish Spp., and white bass from Ellison Creek Reservoir poses an apparent hazard to public health. Therefore, SALG risk assessors recommend that:

1. People should not consume hybrid striped bass and common carp from Ellison Creek Reservoir (Table 10).
2. Women of childbearing age (Women and girls under 50) including pregnant women, women who may become pregnant, and women who are nursing infants and children less than 12 years of age, or who weigh less than 75 pounds should not consume channel catfish, common carp, flathead catfish, hybrid striped bass, spotted gar, sunfish Spp., and white bass from Ellison Creek Reservoir.
3. Women of childbearing age (Women and girls under 50) including pregnant women, women who may become pregnant, and women who are nursing infants and children less than 12 years of age, or who weigh less than 75 pounds may consume up to one four-ounce meal per month of largemouth bass from Ellison Creek Reservoir.
4. Women past childbearing age (Women 50 and older) and males 12 and older may consume up to one eight-ounce meal per month of channel catfish, flathead catfish, spotted gar, or sunfish Spp. from Ellison Creek Reservoir.
5. Women past childbearing age (Women 50 and older) and males 12 and older may consume up to two eight-ounce meal per month of largemouth bass or white bass from Ellison Creek Reservoir.
6. The DSHS advise TPWD to continue not stocking hybrid striped bass (i.e., Palmetto bass) in Ellison Creek Reservoir because hybrid striped bass bioaccumulate significant concentrations of PCBs and PCDDs/PCDFs that pose apparent hazards to public health. The TPWD discontinued stocking of hybrid striped bass following the issuance of the Ellison Creek Reservoir fish consumption advisory in 2005.
7. As resources become available, the DSHS should continue to monitor fish from Ellison Creek Reservoir for changes and establish trends in contaminants of concern or contaminant concentrations that would require a change in consumption advice.

## PUBLIC HEALTH ACTION PLAN

Communication to the public of new and continuing possession bans or consumption advisories, or the removal of either, is essential to effective management of risk from consuming contaminated fish. In fulfillment of the responsibility for communication, the DSHS takes several steps.

- The agency publishes fish consumption advisories and bans in a booklet available to the public through the SALG. To receive the booklet and/or the data, please contact the SALG at 512-834-6757. ${ }^{49}$
- The SALG also posts the most current information about advisories, bans, and the removal of either on the internet at http://www.dshs.state.tx.us/seafood. ${ }^{50}$ The SALG regularly updates this Web site.
- The DSHS also provides the USEPA (http://epa.gov/waterscience/fish/advisories/), the TCEQ (http://www.tceq.state.tx.us), and the TPWD (http://www.tpwd.state.tx.us) with information on all consumption advisories and possession bans. Each year, the TPWD informs the public of consumption advisories and fishing bans on its Web site and in an official downloadable PDF file containing general hunting and fishing regulations available at https://tpwd.texas.gov/publications/pwdpubs/media/cs_bk_k0700_0284.pdf. A booklet containing this information is available at all establishments selling Texas fishing licenses. ${ }^{46}$

Communication to the public of scientific information related to this risk characterization and information for environmental contaminants found in seafood is essential to effective risk management. To achieve this responsibility for communication, the DSHS provides contact information to ask specific questions and/or resources to obtain more information about environmental contaminants in fish.

- Readers may direct questions about the scientific information or recommendations in this risk characterization to the SALG at 512-834-6757 or may find the information at the SALG's Web site (http://www.dshs.state.tx.us/seafood). Secondarily, one may address inquiries to the Environmental and Injury Epidemiology and Toxicology Unit of DSHS (800-588-1248).
- The USEPA's IRIS Web site (http://www.epa.gov/iris/) contains information on environmental contaminants found in food and environmental media.
- The ATSDR, Division of Toxicology (888-42-ATSDR or 888-422-8737 or the ATSDR's Web site (http://www.atsdr.cdc.gov) supplies brief information via ToxFAQs. ${ }^{\text {TM }}$ ToxFAQs $^{\text {TM }}$ are available on the ATSDR Web site in either English or Spanish (http://www.atsdr.cdc.gov/toxfaqs/index.asp). The ATSDR also publishes more in-depth reviews of many toxic substances in its Toxicological Profiles (ToxProfiles ${ }^{\top M}$ ) http://www.atsdr.cdc.gov/toxprofiles/index.asp. To request a copy of the ToxProfiles ${ }^{\top M}$

CD-ROM, PHS, or ToxFAQs ${ }^{\text {TM }}$ call 1-800-CDC-INFO (800-232-4636) or email a request to cdcinfo@cdc.gov.

Figure 1. Ellison Creek Reservoir Map


Figure 2. The relationship between PCB concentration and total length for fish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 3. The relationship between PCB concentration and percent lipids for fish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 4. Length at age for channel catfish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 5. The relationship between PCB concentration and total length for channel catfish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 6. The relationship between PCB concentration and age for channel catfish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 7. Length at age for crappie collected from Ellison Creek Reservoir, Texas, 2014.


Figure 8. Length at age for flathead catfish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 9. Length at age for largemouth bass collected from Ellison Creek Reservoir, Texas, 2014.


Figure 10. The relationship between PCB concentration and percent lipid for largemouth bass collected from Ellison Creek Reservoir, Texas, 2014.


Figure 11. The relationship between PCDD/PCDF concentration and total length for fish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 12. The relationship between PCDD/PCDF concentration and percent lipid for fish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 13. The relationship between PCDD/PCDF concentration and total length for channel catfish collected from Ellison Creek Reservoir, Texas, 2014.


Figure 14. The relationship between PCDD/PCDF concentration and percent lipid for channel catfish collected from Ellison Creek Reservoir, Texas, 2014.


## TABLES

Table 1. Fish samples collected from Ellison Creek Reservoir 2014. Sample number, species, total length, and weight recorded for each sample.

| Sample Number | Species | Total Length |  | Weight |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Millimeters (mm) | Inches (in) | Grams (g) | Pounds (lb) |
| Site 1 Ellison Creek Reservoir at the Dam |  |  |  |  |  |
| ECR1 | Hybrid striped | 622 | 24.5 | 4052 | 8.9 |
| ECR2 | Largemouth bass | 535 | 21.1 | 1971 | 4.3 |
| ECR4 | Flathead catfish | 669 | 26.3 | 4373 | 9.6 |
| ECR5 | Flathead catfish | 506 | 19.9 | 1493 | 3.3 |
| ECR6 | Flathead catfish | 466 | 18.3 | 1215 | 2.7 |
| ECR7 | Channel catfish | 531 | 20.9 | 1665 | 3.7 |
| ECR8 | Channel catfish | 536 | 21.1 | 1523 | 3.4 |
| ECR9 | Channel catfish | 491 | 19.3 | 1064 | 2.3 |
| ECR10 | Channel catfish | 475 | 18.7 | 1109 | 2.4 |
| ECR11 | Channel catfish | 470 | 18.5 | 919 | 2.0 |
| ECR12 | Channel catfish | 408 | 16.1 | 651 | 1.4 |
| ECR13 | Common carp | 656 | 25.8 | 3853 | 8.5 |
| ECR14 | Common carp | 587 | 23.1 | 2702 | 6.0 |
| ECR15 | Largemouth bass | 509 | 20.0 | 1921 | 4.2 |
| ECR16 | Largemouth bass | 481 | 18.9 | 1733 | 3.8 |
| ECR17 | Largemouth bass | 443 | 17.4 | 1366 | 3.0 |
| ECR18 | Largemouth bass | 445 | 17.5 | 1251 | 2.8 |
| ECR19 | Largemouth bass | 449 | 17.7 | 1124 | 2.5 |
| ECR20 | Largemouth bass | 410 | 16.1 | 1014 | 2.2 |
| ECR22 | Largemouth bass | 407 | 16.0 | 1055 | 2.3 |
| ECR23 | Channel catfish | 400 | 15.7 | 676 | 1.5 |
| ECR24 | Channel catfish | 376 | 14.8 | 508 | 1.1 |
| ECR86 | Black crappie | 281 | 11.1 | 335 | 0.7 |
| ECR87 | Channel catfish | 470 | 18.5 | 952 | 2.1 |
| ECR88 | Channel catfish | 470 | 18.5 | 1042 | 2.3 |
| ECR89 | Flathead catfish | 701 | 27.6 | 4892 | 10.8 |
| ECR90 | Flathead catfish | 900 | 35.4 | 8800 | 19.4 |
| ECR93 | Flathead catfish | 770 | 30.3 | 5727 | 12.6 |
| ECR112 | Channel catfish | 483 | 19.0 | 1424 | 3.1 |
| Site 2 Ellison Creek Reservoir AEP Intake |  |  |  |  |  |
| ECR25 | Largemouth bass | 419 | 16.5 | 1052 | 2.3 |
| ECR26 | Largemouth bass | 385 | 15.2 | 843 | 1.9 |
| ECR27 | Largemouth bass | 440 | 17.3 | 1266 | 2.8 |
| ECR28 | Largemouth bass | 400 | 15.7 | 929 | 2.0 |
| ECR29 | Common carp | 696 | 27.4 | 4160 | 9.2 |
| ECR30 | Common carp | 655 | 25.8 | 4443 | 9.8 |

Table 1. cont. Fish samples collected from Ellison Creek Reservoir 2014. Sample number, species, total length, and weight recorded for each sample.

| Sample Number | Species | Total Length |  | Weight |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Millimeters (mm) | Inches (in) | Grams (g) | Pounds (lb) |
| Site 2 Cont. Ellison Creek Reservoir AEP Intake |  |  |  |  |  |
| ECR31 | Common carp | 651 | 25.6 | 3966 | 8.7 |
| ECR32 | Flathead catfish | 744 | 29.3 | 5492 | 12.1 |
| ECR33 | Channel catfish | 557 | 21.9 | 2463 | 5.4 |
| ECR34 | Channel catfish | 502 | 19.8 | 1614 | 3.6 |
| ECR35 | Channel catfish | 557 | 21.9 | 2071 | 4.6 |
| ECR36 | Channel catfish | 526 | 20.7 | 1771 | 3.9 |
| ECR37 | Channel catfish | 521 | 20.5 | 1568 | 3.5 |
| ECR38 | Channel catfish | 410 | 16.1 | 736 | 1.6 |
| ECR39 | Largemouth bass | 497 | 19.6 | 2046 | 4.5 |
| ECR40 | Hybrid striped | 621 | 24.4 | 3901 | 8.6 |
| ECR49 | Channel catfish | 412 | 16.2 | 625 | 1.4 |
| ECR50 | Channel catfish | 395 | 15.6 | 571 | 1.3 |
| ECR51 | Common carp | 739 | 29.1 | 5760 | 12.7 |
| ECR52 | Largemouth bass | 480 | 18.9 | 1847 | 4.1 |
| ECR53 | Channel catfish | 542 | 21.3 | 1566 | 3.5 |
| ECR85 | White bass | 352 | 13.9 | 706 | 1.6 |
| Site 3 Ellison Creek Reservoir at Barnes Creek |  |  |  |  |  |
| ECR41 | Flathead catfish | 843 | 33.2 | 8624 | 19.0 |
| ECR42 | Channel catfish | 629 | 24.8 | 3512 | 7.7 |
| ECR43 | Channel catfish | 450 | 17.7 | 866 | 1.9 |
| ECR44 | Channel catfish | 385 | 15.2 | 458 | 1.0 |
| ECR45 | Largemouth bass | 529 | 20.8 | 2004 | 4.4 |
| ECR46 | Common carp | 626 | 24.6 | 3136 | 6.9 |
| ECR47 | Common carp | 664 | 26.1 | 4083 | 9.0 |
| ECR48 | Common carp | 672 | 26.5 | 4569 | 10.1 |
| ECR91 | Spotted gar | 648 | 25.5 | 1178 | 2.6 |
| ECR92 | Spotted gar | 691 | 27.2 | 1150 | 2.5 |
| ECR94 | Green sunfish | 226 | 8.9 | 265 | 0.6 |
| ECR95 | Black crappie | 260 | 10.2 | 265 | 0.6 |
| ECR96 | Black crappie | 275 | 10.8 | 322 | 0.7 |
| ECR97 | Bowfin | 525 | 20.7 | 1624 | 3.6 |
| ECR98 | Bowfin | 600 | 23.6 | 2338 | 5.2 |
| ECR99 | Channel catfish | 541 | 21.3 | 1621 | 3.6 |
| ECR100 | Channel catfish | 440 | 17.3 | 1004 | 2.2 |
| ECR101 | Channel catfish | 446 | 17.6 | 1118 | 2.5 |
| ECR102 | Channel catfish | 453 | 17.8 | 918 | 2.0 |
| ECR103 | Largemouth bass | 520 | 20.5 | 2319 | 5.1 |

Table 1. cont. Fish samples collected from Ellison Creek Reservoir 2014. Sample number, species, total length, and weight recorded for each sample.

| Sample Number | Species | Total Length |  | Weight |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Millimeters (mm) | Inches (in) | Grams (g) | Pounds (lb) |
| Site 3 cont. Ellison Creek Reservoir at Barnes Creek |  |  |  |  |  |
| ECR104 | Largemouth bass | 543 | 21.4 | 2568 | 5.7 |
| ECR105 | Largemouth bass | 495 | 19.5 | 1735 | 3.8 |
| ECR106 | Largemouth bass | 470 | 18.5 | 1513 | 3.3 |
| ECR107 | Largemouth bass | 471 | 18.5 | 1334 | 2.9 |
| ECR108 | Largemouth bass | 425 | 16.7 | 1312 | 2.9 |
| ECR109 | Largemouth bass | 441 | 17.4 | 1294 | 2.9 |
| ECR111 | Spotted gar | 715 | 28.1 | 1751 | 3.9 |
| Site 4 Ellison Creek Reservoir Upper Reservoir |  |  |  |  |  |
| ECR54 | Channel catfish | 631 | 24.8 | 3215 | 7.1 |
| ECR55 | Channel catfish | 475 | 18.7 | 1093 | 2.4 |
| ECR56 | Channel catfish | 559 | 22.0 | 1598 | 3.5 |
| ECR57 | Channel catfish | 625 | 24.6 | 3168 | 7.0 |
| ECR59 | Channel catfish | 529 | 20.8 | 1595 | 3.5 |
| ECR60 | White bass | 393 | 15.5 | 725 | 1.6 |
| ECR61 | Largemouth bass | 477 | 18.8 | 1514 | 3.3 |
| ECR62 | Common carp | 679 | 26.7 | 4713 | 10.4 |
| ECR63 | Common carp | 630 | 24.8 | 3859 | 8.5 |
| ECR64 | Common carp | 689 | 27.1 | 4264 | 9.4 |
| ECR65 | Common carp | 635 | 25.0 | 3651 | 8.0 |
| ECR66 | Spotted gar | 686 | 27.0 | 1198 | 2.6 |
| ECR67 | Bowfin | 701 | 27.6 | 3077 | 6.8 |
| ECR68 | White crappie | 346 | 13.6 | 545 | 1.2 |
| ECR69 | White crappie | 322 | 12.7 | 478 | 1.1 |
| ECR70 | White crappie | 348 | 13.7 | 495 | 1.1 |
| ECR71 | Black crappie | 272 | 10.7 | 282 | 0.6 |
| ECR72 | Largemouth bass | 580 | 22.8 | 3080 | 6.8 |
| ECR73 | Largemouth bass | 476 | 18.7 | 1580 | 3.5 |
| ECR74 | Largemouth bass | 551 | 21.7 | 2726 | 6.0 |
| ECR75 | Largemouth bass | 513 | 20.2 | 1947 | 4.3 |
| ECR76 | Largemouth bass | 426 | 16.8 | 1080 | 2.4 |
| ECR77 | Largemouth bass | 391 | 15.4 | 897 | 2.0 |
| ECR78 | Largemouth bass | 373 | 14.7 | 641 | 1.4 |
| ECR79 | Channel catfish | 454 | 17.9 | 982 | 2.2 |
| ECR80 | Channel catfish | 491 | 19.3 | 1150 | 2.5 |
| ECR81 | Channel catfish | 441 | 17.4 | 761 | 1.7 |
| ECR84 | Redbreast sunfish | 201 | 7.9 | 141 | 0.3 |
| ECR113 | Redbreast sunfish | 226 | 8.9 | 249 | 0.5 |
| ECR114 | Bluegill | 182 | 7.2 | 135 | 0.3 |

Table 2.1. Arsenic ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by species, 2014.

| Species | $\begin{array}{c}\text { Number } \\ \text { Detected/ } \\ \text { Number Tested }\end{array}$ | $\begin{array}{c}\text { Total Arsenic } \\ \text { Mean } \pm \text { S.D. } \\ \text { (Min-Max) }\end{array}$ | $\begin{array}{c}\text { Inorganic } \\ \text { Arsenic } \\ \text { Mean }\end{array}$ | $\begin{array}{c}\text { HAC Value } \\ \text { (nonca) and } \\ \text { HAC Value } \\ \text { (ca; mg/kg) }\end{array}$ | Basis for Comparison Value |
| :--- | :---: | :---: | :---: | :---: | :---: |$]$|  |
| :---: |
| Black crappie |
| Bowfin |

[^5]

| Species | Number Detected/ <br> Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value <br> (nonca; mg/kg) |
| :--- | :---: | :---: | :---: |
| Black crappie | $4 / 4$ |  | Basis for Comparison Value |



| Species | Number Detected/ Number Tested | $\begin{gathered} \text { Mean } \pm \text { S.D. } \\ \text { (Min-Max) } \end{gathered}$ | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Black crappie | 4/4 | BDL | 334 | Based on the Tolerable Upper Intake Level (UL) $-0.143 \mathrm{mg} / \mathrm{kg}$-day ${ }^{\mathrm{j}}$ |
| Bowfin | 3/3 | $\begin{aligned} & 1.038 \pm 1.451 \\ & \text { (BDL-2.714) } \end{aligned}$ |  |  |
| Channel catfish | 33/33 | $\begin{aligned} & 0.228 \pm 0.160 \\ & \text { (BDL-1.117) } \end{aligned}$ |  |  |
| Common carp | 13/13 | $\begin{aligned} & 0.535 \pm 0.313 \\ & (B D L-1.033) \end{aligned}$ |  |  |
| Flathead catfish | 8/8 | BDL |  |  |
| Hybrid striped bass | 2/2 | BDL |  |  |
| Largemouth bass | 30/30 | BDL |  |  |
| Spotted gar | 4/4 | BDL |  |  |
| White bass | 2/2 | BDL |  |  |
| White crappie | 3/3 | BDL |  |  |
| Crappie Spp. (black and white) | 7/7 | BDL |  |  |
| All fish combined | 100/100 | $\begin{aligned} & 0.278 \pm 0.305 \\ & (B D L-2.714) \end{aligned}$ |  |  |

[^6]Table 2.4. Lead ( $\mathrm{mg} / \mathrm{kg} \mathrm{)} \mathrm{in} \mathrm{fish} \mathrm{collected} \mathrm{from} \mathrm{Ellison} \mathrm{Creek} \mathrm{Reservoir} \mathrm{by} \mathrm{species}, \mathrm{2014}$.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. (Min-Max) | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Black crappie | 4/4 | $\begin{aligned} & \hline 0.522 \pm 0.216 \\ & (B D L-0.658) \end{aligned}$ | N/A | N/A |
| Bowfin | 3/3 | BDL |  |  |
| Channel catfish | 32/33 | $\begin{gathered} \hline 0.206 \pm 0.035 \\ (N D-0.401) \\ \hline \end{gathered}$ |  |  |
| Common carp | 11/13 | ND-BDL |  |  |
| Flathead catfish | 8/8 | BDL |  |  |
| Hybrid striped bass | 2/2 | $\begin{aligned} & 0.472 \pm 0.385 \\ & \text { (BDL-0.744) } \\ & \hline \end{aligned}$ |  |  |
| Largemouth bass | 30/30 | $\begin{aligned} & 0.296 \pm 0.171 \\ & (B D L-0.753) \\ & \hline \end{aligned}$ |  |  |
| Spotted gar | 4/4 | $\begin{aligned} & 0.270 \pm 0.141 \\ & (B D L-0.482) \\ & \hline \end{aligned}$ |  |  |
| White bass | 2/2 | $\begin{aligned} & \hline 0.605 \pm 0.573 \\ & \text { (BDL-1.011) } \\ & \hline \end{aligned}$ |  |  |
| White crappie | 3/3 | BDL |  |  |
| Crappie Spp. (black and white) | 7/7 | $\begin{aligned} & \hline 0.384 \pm 0.230 \\ & \text { (BDL-0.658) } \\ & \hline \end{aligned}$ |  |  |
| Temperate bass | 4/4 | $\begin{aligned} & 0.539 \pm 0.406 \\ & (B D L-1.011) \\ & \hline \end{aligned}$ |  |  |
| All fish combined | 97/100 | $\begin{gathered} \hline 0.258 \pm 0.154 \\ (N D-1.011) \end{gathered}$ |  |  |

Table 2.5. Selenium ( $\mathrm{mg} / \mathrm{kg} \mathrm{)} \mathrm{in} \mathrm{fish} \mathrm{collected} \mathrm{from} \mathrm{Ellison} \mathrm{Creek} \mathrm{Reservoir} \mathrm{by} \mathrm{species}, \mathrm{2014}$.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. (Min-Max) | HAC Value (nonca; $\mathrm{mg} / \mathrm{kg}$ ) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Black crappie | 4/4 | $\begin{aligned} & \hline 0.178 \pm 0.256 \\ & \text { (BDL-0.563) } \end{aligned}$ | 6 | EPA Chronic Oral RfD - $0.005 \mathrm{mg} / \mathrm{kg}$-day ATSDR Chronic Oral MRL $-0.005 \mathrm{mg} / \mathrm{kg}$-day UL: $0.400 \mathrm{mg} /$ day ( $0.005 \mathrm{mg} / \mathrm{kg}$-day) <br> RfD or MRL/2 - $0.005 \mathrm{mg} / \mathrm{kg}-$ day $/ 2=0.0025$ $\mathrm{mg} / \mathrm{kg}$-day) ${ }^{\mathrm{k}, 51}$ |
| Bowfin | 3/3 | $\begin{aligned} & \hline 0.207 \pm 0.157 \\ & \text { (BDL-0.363) } \\ & \hline \end{aligned}$ |  |  |
| Channel catfish | 33/33 | $\begin{aligned} & \hline 0.425 \pm 1.021 \\ & (B D L-4.477) \end{aligned}$ |  |  |
| Common carp | 13/13 | $\begin{aligned} & \hline 0.506 \pm 0.157 \\ & \text { (BDL-0.661) } \\ & \hline \end{aligned}$ |  |  |
| Flathead catfish | 8/8 | $\begin{aligned} & \hline 0.148 \pm 0.145 \\ & \text { (BDL-0.470) } \\ & \hline \end{aligned}$ |  |  |
| Hybrid striped bass | 2/2 | $\begin{gathered} \hline 0.332 \pm 0.284 \\ (0.131-0.533) \\ \hline \end{gathered}$ |  |  |
| Largemouth bass | 30/30 | $\begin{aligned} & 0.567 \pm 0.988 \\ & \text { (BDL-3.528) } \end{aligned}$ |  |  |
| Spotted gar | 4/4 | $\begin{aligned} & \hline 0.165 \pm 0.184 \\ & \text { (BDL-0.436) } \\ & \hline \end{aligned}$ |  |  |
| White bass | 2/2 | $\begin{aligned} & \hline 0.433 \pm 0.542 \\ & (B D L-0.816) \end{aligned}$ |  |  |
| White crappie | 3/3 | $\begin{gathered} \hline 0.569 \pm 0.089 \\ (0.487-0.663) \\ \hline \end{gathered}$ |  |  |
| Crappie Spp. (black and white) | 7/7 | $\begin{aligned} & \hline 0.346 \pm 0.281 \\ & \text { (BDL-0.663) } \\ & \hline \end{aligned}$ |  |  |
| Temperate bass | 4/4 | $\begin{aligned} & \hline 0.382 \pm 0.358 \\ & \text { (BDL-0.816) } \\ & \hline \end{aligned}$ |  |  |
| All fish combined | 100/100 | $\begin{aligned} & 0.429 \pm 0.797 \\ & \text { (BDL-4.477) } \\ & \hline \end{aligned}$ |  |  |

[^7]Table 2.6. Zinc ( $\mathrm{mg} / \mathrm{kg} \mathrm{)} \mathrm{in} \mathrm{fish} \mathrm{collected} \mathrm{from} \mathrm{Ellison} \mathrm{Creek} \mathrm{Reservoir} \mathrm{by} \mathrm{species}, \mathrm{2014}$.

| Species | Number Detected/ Number Tested | $\begin{gathered} \text { Mean } \pm \text { S.D. } \\ \text { (Min-Max) } \end{gathered}$ | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Black crappie | 4/4 | $\begin{gathered} \hline 3.952 \pm 0.467 \\ (3.371-4.505) \\ \hline \end{gathered}$ | 700 | EPA Chronic Oral RfD - $0.3 \mathrm{mg} / \mathrm{kg}$-day |
| Bowfin | 3/3 | $\begin{gathered} \hline 2.848 \pm 0.622 \\ (2.357-3.548) \\ \hline \end{gathered}$ |  |  |
| Channel catfish | 33/33 | $\begin{gathered} \hline 4.694 \pm 1.332 \\ (2.764-9.781) \\ \hline \end{gathered}$ |  |  |
| Common carp | 13/13 | $\begin{gathered} \hline 5.842 \pm 2.351 \\ (1.260-11.452) \\ \hline \end{gathered}$ |  |  |
| Flathead catfish | 8/8 | $\begin{gathered} \hline 3.477 \pm 0.433 \\ (2.718-3.971) \\ \hline \end{gathered}$ |  |  |
| Hybrid striped bass | 2/2 | $\begin{gathered} 2.856 \pm 0.120 \\ (2.771-2.941) \end{gathered}$ |  |  |
| Largemouth bass | 30/30 | $\begin{gathered} 3.356 \pm 0.804 \\ (1.307-5.337) \end{gathered}$ |  |  |
| Spotted gar | 4/4 | $\begin{gathered} 2.622 \pm 0.330 \\ (2.334-2.938) \\ \hline \end{gathered}$ |  |  |
| White bass | 2/2 | $\begin{gathered} 3.346 \pm 0.337 \\ (3.108-3.585) \\ \hline \end{gathered}$ |  |  |
| White crappie | 3/3 | $\begin{gathered} 4.113 \pm 0.645 \\ (3.647-4.849) \\ \hline \end{gathered}$ |  |  |
| Crappie Spp. (black and white) | 7/7 | $\begin{gathered} 4.021 \pm 0.505 \\ (3.371-4.849) \\ \hline \end{gathered}$ |  |  |
| Temperate bass | 4/4 | $\begin{gathered} 3.101 \pm 0.351 \\ (2.771-3.585) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 100/100 | $\begin{gathered} \hline 4.122 \pm 1.531 \\ (1.260-11.452) \\ \hline \end{gathered}$ |  |  |

Table 2.7. Mercury ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by sample site, 2014.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. (Min-Max) | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Site 1 Ellison Creek Reservoir at the Dam |  |  |  |  |
| Black crappie | 1/1 | 0.031 | 0.7 | ATSDR Chronic Oral MRL for Methylmercury$-0.0003 \mathrm{mg} / \mathrm{kg}-\text { day }$ |
| Channel catfish | 11/11 | $\begin{gathered} \hline 0.042 \pm 0.063 \\ (0.009-0.216) \\ \hline \end{gathered}$ |  |  |
| Common carp | 2/2 | $\begin{gathered} \hline 0.031 \pm 0.006 \\ (0.027-0.035) \\ \hline \end{gathered}$ |  |  |
| Flathead catfish | 6/6 | $\begin{gathered} 0.058 \pm 0.031 \\ (0.029-0.106) \\ \hline \end{gathered}$ |  |  |
| Hybrid striped bass | 1/1 | 0.124 |  |  |
| Largemouth bass | 8/8 | $\begin{gathered} \hline 0.143 \pm 0.114 \\ (0.068-0.416) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 29/29 | $\begin{gathered} \hline 0.075 \pm 0.083 \\ (0.009-0.416) \\ \hline \end{gathered}$ |  |  |

Site 2 Ellison Creek Reservoir near AEP Intake

| Channel catfish | 9/9 | $\begin{gathered} 0.022 \pm 0.011 \\ (0.010-0.042) \end{gathered}$ | 0.7 | ATSDR Chronic Oral MRL for Methylmercury$-0.0003 \mathrm{mg} / \mathrm{kg}-\text { day }$ |
| :---: | :---: | :---: | :---: | :---: |
| Common carp | 4/4 | $\begin{gathered} 0.026 \pm 0.011 \\ (0.012-0.035) \\ \hline \end{gathered}$ |  |  |
| Flathead catfish | 1/1 | 0.044 |  |  |
| Hybrid striped bass | 1/1 | 0.213 |  |  |
| Largemouth bass | 6/6 | $\begin{gathered} 0.076 \pm 0.018 \\ (0.058-0.108) \\ \hline \end{gathered}$ |  |  |
| White bass | 1/1 | 0.122 |  |  |
| All fish combined | 22/22 | $\begin{gathered} \hline 0.052 \pm 0.048 \\ (0.010-0.213) \\ \hline \end{gathered}$ |  |  |

Site 3 Ellison Creek Reservoir at Barnes Creek

| Black crappie | $2 / 2$ | $0.032 \pm 0.009$ <br> $(0.025-0.038)$ |  |
| :--- | :---: | :---: | :---: |
| Bowfin | $2 / 2$ | $0.364 \pm 0.231$ <br> $(0.201-0.527)$ |  |
| Channel catfish | $7 / 7$ | $0.032 \pm 0.031$ <br> $(0.009-0.095)$ |  |
| Common carp | $3 / 3$ | $0.038 \pm 0.007$ <br> $(0.031-0.045)$ |  |
| Flathead catfish | $1 / 1$ | 0.7 |  |
| Green sunfish | $1 / 1$ | 0.101 |  |
| Largemouth bass | $8 / 8$ | $0.155 \pm 0.047$ <br> $(0.094-0.229)$ |  |
| Spotted gar | $3 / 3$ | $0.101 \pm 0.066$ <br> $(0.051-0.176)$ |  |
| All fish combined | $27 / 27$ | $0.104 \pm 0.108$ <br> $(0.009-0.527)$ |  |

Table 2.8. Mercury ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by sample site, 2014.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Site 4 Ellison Creek Reservoir Upper Reservoir |  |  |  |  |
| Black crappie | 1/1 | 0.027 | 0.7 | ATSDR Chronic Oral MRL for Methylmercury $-0.0003 \mathrm{mg} / \mathrm{kg}$-day |
| Bluegill | 1/1 | 0.012 |  |  |
| Bowfin | 1/1 | 0.417 |  |  |
| Channel catfish | 8/8 | $\begin{gathered} \hline 0.036 \pm 0.020 \\ (0.009-0.071) \\ \hline \end{gathered}$ |  |  |
| Common carp | 4/4 | $\begin{gathered} 0.059 \pm 0.015 \\ (0.043-0.079) \\ \hline \end{gathered}$ |  |  |
| Largemouth bass | 8/8 | $\begin{gathered} 0.113 \pm 0.053 \\ (0.066-0.231) \\ \hline \end{gathered}$ |  |  |
| Redbreast sunfish | 2/2 | $\begin{gathered} \hline 0.014 \pm 0.001 \\ (0.013-0.014) \\ \hline \end{gathered}$ |  |  |
| Spotted gar | 1/1 | 0.080 |  |  |
| White bass | 1/1 | 0.081 |  |  |
| White crappie | 3/3 | $\begin{gathered} \hline 0.056 \pm 0.017 \\ (0.039-0.072) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 30/30 | $\begin{gathered} 0.075 \pm 0.079 \\ (0.009-0.417) \\ \hline \end{gathered}$ |  |  |

Table 2.9. Mercury ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by species, 2014.

| Species |  |  |
| :--- | :---: | :---: | :---: |

Table 3. Pesticides ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by species, 2014.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value (nonca) and HAC Value (ca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Chlordane (sum) |  |  |  |  |
| Channel catfish | 8/8 | $\begin{gathered} \hline 0.0110 \pm 0.0045 \\ (0.0076-0.0216) \\ \hline \end{gathered}$ | 1.1671.556 | EPA Chronic Oral RfD $-0.0005 \mathrm{mg} / \mathrm{kg}$-day <br> EPA Oral Slope Factor -0.35 per $\mathrm{mg} / \mathrm{kg}$-day |
| Largemouth bass | 8/8 | $\begin{gathered} \hline 0.0026 \pm 0.0019 \\ (0.0010-0.0067) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 16/16 | $\begin{gathered} 0.0068 \pm 0.0055 \\ (0.0010-0.0216) \end{gathered}$ |  |  |
| 4,4'-DDE |  |  |  |  |
| Channel catfish | 8/8 | $\begin{gathered} \hline 0.0061 \pm 0.0027 \\ (0.0040-0.0117) \end{gathered}$ | 1.1671.601 | EPA Chronic Oral RfD for DDT - 5.0E-4 $\mathrm{mg} / \mathrm{kg}$-day <br> EPA Oral Slope Factor for DDT- 3.4E-1 per $\mathrm{mg} / \mathrm{kg}$-day |
| Largemouth bass | 8/8 | $\begin{gathered} 0.0020 \pm 0.0014 \\ (0.0007-0.0047) \end{gathered}$ |  |  |
| All fish combined | 16/16 | $\begin{gathered} 0.0041 \pm 0.0029 \\ (0.0007-0.0117) \end{gathered}$ |  |  |
| Pentachloroanisole |  |  |  |  |
| Channel catfish | 8/8 | $\begin{gathered} 0.0013 \pm 0.0009 \\ (0.0004-0.0034) \end{gathered}$ | N/A | N/A |
| Largemouth bass | 0/8 | ND |  |  |
| All fish combined | 8/16 | $\begin{gathered} 0.0007 \pm 0.0009 \\ \text { (ND-0.0034) } \\ \hline \end{gathered}$ |  |  |

Table 4.1. PCBs ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by sample site, 2014.

| Species | Number Detected/ <br> Number Tested |  | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value <br> (nonca; mg/kg) |
| :--- | :---: | :---: | :---: | :---: |
| Site 1 Ellison Creek Reservoir at the Dam | Basis for Comparison Value |  |  |  |

Site 2 Ellison Creek Reservoir near AEP Intake

| Channel catfish | $9 / 9$ | $\mathbf{0 . 1 1 3} \pm 0.039$ <br> $(\mathbf{0 . 0 7 4 - 0 . 1 9 8 )}$ |  |  |
| :--- | :---: | :---: | :---: | :---: |
| Common carp | $4 / 4$ | $\mathbf{0 . 2 2 8} \pm 0.048$ <br> $(0.160-0.273)$ |  |  |
| Flathead catfish | $1 / 1$ | $\mathbf{0 . 1 2 9}$ | 0.047 | 0.272 |

[^8]Table 4.2. PCBs ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from Ellison Creek Reservoir by sample site, 2014.

| Species | Number Detected/ Number Tested | $\begin{gathered} \text { Mean } \pm \text { S.D. } \\ \text { (Min-Max) } \end{gathered}$ | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Site 3 Ellison Creek Reservoir at Barnes Creek |  |  |  |  |
| Black crappie | 2/2 | $\begin{gathered} 0.010 \pm 0.001 \\ (0.009-0.011) \\ \hline \end{gathered}$ | 0.047 | EPA Chronic Oral RfD for Aroclor 1254 $0.00002 \mathrm{mg} / \mathrm{kg}$-day |
| Bowfin | 2/2 | $\begin{gathered} 0.017 \pm 0.009 \\ (0.010-0.023) \\ \hline \end{gathered}$ |  |  |
| Channel catfish | 7/7 | $\begin{aligned} & 0.140^{m} \pm 0.122 \\ & (0.039-0.368) \\ & \hline \end{aligned}$ |  |  |
| Common carp | 3/3 | $\begin{gathered} \hline 0.172 \pm 0.079 \\ (0.098-0.256) \\ \hline \end{gathered}$ |  |  |
| Flathead catfish | 1/1 | 0.065 |  |  |
| Green sunfish | 1/1 | 0.012 | 0.272 | EPA Slope Factor - 2.0 per mg/kg-day |
| Largemouth bass | 8/8 | $\begin{gathered} 0.088 \pm 0.061 \\ (0.034-0.227) \end{gathered}$ |  |  |
| Spotted gar | 3/3 | $\begin{gathered} 0.117 \pm 0.073 \\ (0.036-0.179) \end{gathered}$ |  |  |
| All fish combined | 27/27 | $\begin{gathered} 0.099 \pm 0.089 \\ (0.009-0.368) \\ \hline \end{gathered}$ |  |  |

Site 4 Ellison Creek Reservoir Upper Reservoir

| Black crappie | 1/1 | 0.026 | 0.047 | $\begin{aligned} & \text { EPA Chronic Oral RfD for Aroclor } 1254 \text { - } \\ & 0.00002 \mathrm{mg} / \mathrm{kg} \text {-day } \\ & \text { EPA Slope Factor }-2.0 \text { per } \mathrm{mg} / \mathrm{kg} \text {-day } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| Bluegill | 1/1 | 0.162 |  |  |
| Bowfin | 1/1 | 0.048 |  |  |
| Channel catfish | 8/8 | $\begin{gathered} \hline 0.132 \pm 0.116 \\ (0.054-0.408) \\ \hline \end{gathered}$ |  |  |
| Common carp | 4/4 | $\begin{gathered} \hline 0.223 \pm 0.141 \\ (0.123-0.423) \\ \hline \end{gathered}$ |  |  |
| Largemouth bass | 8/8 | $\begin{gathered} \hline 0.054 \pm 0.022 \\ (0.018-0.096) \end{gathered}$ |  |  |
| Redbreast sunfish | 1/1 | 0.136 | 0.272 |  |
| Spotted gar | 1/1 | 0.092 |  |  |
| White bass | 1/1 | 0.088 |  |  |
| White crappie | 3/3 | $\begin{gathered} 0.029 \pm 0.018 \\ (0.013-0.048) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 29/29 | $\begin{gathered} 0.104 \pm 0.099 \\ (0.013-0.423) \end{gathered}$ |  |  |

[^9]Table 4.3. PCBs (mg/kg) in fish collected from Ellison Creek Reservoir by species, 2014.

| Species |  |  |
| :--- | :---: | :---: | :---: |

[^10]Table 5.1. PCDDs/PCDFs toxicity equivalent (TEQ) concentrations ( $\mathrm{pg} / \mathrm{g}$ ) in fish collected from the Ellison Creek Reservoir by sample site, 2014.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value (nonca; pg/g) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Site 1 Ellison Creek Reservoir at the Dam |  |  |  |  |
| Black crappie | 1/1 | 0.182 | 2.33 | ATSDR Chronic Oral MRL for 2,3,7,8 - TCDD $-1.0 \times 10-9 \mathrm{mg} / \mathrm{kg}-\text { day }$ <br> EPA Slope Factor $-1.56 \times 105$ per mg/kgday |
| Channel catfish | 10/10 | $\begin{gathered} 0.545 \pm 0.306 \\ (0.142-0.950) \\ \hline \end{gathered}$ |  |  |
| Common carp | 2/2 | $\begin{aligned} & \hline 2.908^{\circ} \pm 0.095 \\ & (2.841-2.975) \\ & \hline \end{aligned}$ |  |  |
| Flathead catfish | 6/6 | $\begin{gathered} \hline 0.939 \pm 0.461 \\ (0.540-1.689) \\ \hline \end{gathered}$ |  |  |
| Hybrid striped bass | 1/1 | 5.928 | 3.49 |  |
| Largemouth bass | 8/8 | $\begin{gathered} 0.370 \pm 0.146 \\ (0.100-0.513) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 28/28 | $\begin{gathered} \hline 0.928 \pm 1.208 \\ (0.100-5.928) \\ \hline \end{gathered}$ |  |  |

Site 2 Ellison Creek Reservoir near AEP Intake

| Channel catfish | 9/9 | $\begin{gathered} 0.948 \pm 0.497 \\ (0.379-1.816) \end{gathered}$ | 2.33 | ATSDR Chronic Oral MRL for 2,3,7,8 - TCDD $-1.0 \times 10-9 \mathrm{mg} / \mathrm{kg}$-day |
| :---: | :---: | :---: | :---: | :---: |
| Common carp | 4/4 | $\begin{gathered} \hline 1.241 \pm 0.290 \\ (0.934-1.521) \\ \hline \end{gathered}$ |  |  |
| Flathead catfish | 1/1 | 0.838 |  |  |
| Hybrid striped bass | 1/1 | 1.678 |  |  |
| Largemouth bass | 5/5 | $\begin{gathered} 0.451 \pm 0.112 \\ (0.308-0.573) \end{gathered}$ | 3.49 | EPA Slope Factor $-1.56 \times 105$ per $\mathrm{mg} / \mathrm{kg}-~$day |
| White bass | 1/1 | 0.769 |  |  |
| All fish combined | 21/21 | $\begin{gathered} \hline 0.906 \pm 0.469 \\ (0.308-1.816) \end{gathered}$ |  |  |

[^11]Table 5.2. PCDDs/PCDFs toxicity equivalent (TEQ) concentrations ( $\mathrm{pg} / \mathrm{g}$ ) in fish collected from the Ellison Creek Reservoir by sample site, 2014.

| Species | Number Detected/ <br> Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value <br> (nonca; pg/g) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |

Site 3 Ellison Creek Reservoir at Barnes Creek

| Black crappie | 1/2 | $\begin{gathered} 0.001 \pm 0.001 \\ (N D-0.002) \\ \hline \end{gathered}$ | 2.33 | ATSDR Chronic Oral MRL for 2,3,7,8 - TCDD$-1.0 \times 10-9 \mathrm{mg} / \mathrm{kg}-\text { day }$ |
| :---: | :---: | :---: | :---: | :---: |
| Bowfin | 2/2 | $\begin{gathered} 0.053 \pm 0.032 \\ (0.031-0.076) \\ \hline \end{gathered}$ |  |  |
| Channel catfish | 6/6 | $\begin{gathered} 1.226 \pm 1.416 \\ \left(0.196-3.338^{\mathrm{p}}\right) \end{gathered}$ |  |  |
| Common carp | 3/3 | $\begin{gathered} \hline 0.942 \pm 0.420 \\ (0.575-1.401) \\ \hline \end{gathered}$ |  |  |
| Flathead catfish | 1/1 | 0.260 | 3.49 | $\begin{gathered} \text { EPA Slope Factor }-1.56 \times 105 \text { per mg/kg- } \\ \text { day } \end{gathered}$ |
| Largemouth bass | 7/8 | $\begin{gathered} 0.518 \pm 0.412 \\ (N D-1.247) \end{gathered}$ |  |  |
| Spotted gar | 3/3 | $\begin{gathered} \hline 0.960 \pm 0.621 \\ (0.253-1.416) \\ \hline \end{gathered}$ |  |  |
| All fish combined | 23/25 | $\begin{gathered} 0.703 \pm 0.833 \\ (N D-3.338) \end{gathered}$ |  |  |

Site 4 Ellison Creek Reservoir Upper Reservoir

| Black crappie | 1/1 | 0.283 | 2.33 | $\begin{aligned} & \text { ATSDR Chronic Oral MRL for 2,3,7,8-TCDD } \\ & \qquad-1.0 \times 10-9 \mathrm{mg} / \mathrm{kg} \text {-day } \\ & \text { EPA Slope Factor }-1.56 \times 105 \text { per } \mathrm{mg} / \mathrm{kg}- \\ & \text { day } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| Bowfin | 1/1 | 0.541 |  |  |
| Channel catfish | 8/8 | $\begin{gathered} 2.789 \pm 3.328 \\ (0.659-10.400) \end{gathered}$ |  |  |
| Common carp | 4/4 | $\begin{gathered} \hline 1.717 \pm 1.181 \\ (0.774-3.241) \\ \hline \end{gathered}$ |  |  |
| Largemouth bass | 7/7 | $\begin{gathered} 0.886 \pm 0.652 \\ (0.077-1.927) \\ \hline \end{gathered}$ |  |  |
| Spotted gar | 1/1 | 0.995 | 3.49 |  |
| White bass | 1/1 | 1.045 |  |  |
| White crappie | 3/3 | $\begin{gathered} 0.169 \pm 0.109 \\ (0.091-0.294) \end{gathered}$ |  |  |
| All fish combined | 26/26 | $\begin{gathered} \hline 1.490 \pm 2.084 \\ (0.077-10.400) \end{gathered}$ |  |  |

[^12]Table 5.3. PCDDs/PCDFs toxicity equivalent (TEQ) concentrations ( $\mathrm{pg} / \mathrm{g}$ ) in fish collected from the Ellison Creek Reservoir by species, 2014.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value (nonca; pg/g) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Black crappie | 3/4 | $\begin{gathered} \hline 0.117 \pm 0.140 \\ (N D-0.283) \\ \hline \end{gathered}$ | 2.33 | ATSDR Chronic Oral MRL for 2,3,7,8 - TCDD $-1.0 \times 10-9 \mathrm{mg} / \mathrm{kg}-\text { day }$ <br> EPA Slope Factor - $1.56 \times 105$ per mg/kg- <br> day |
| Bowfin | 3/3 | $\begin{gathered} 0.216 \pm 0.282 \\ (0.031-0.541) \\ \hline \end{gathered}$ |  |  |
| Channel catfish | 33/33 | $\begin{gathered} 1.323 \pm 1.895 \\ \left(0.142-10.400^{q}\right) \\ \hline \end{gathered}$ |  |  |
| Common carp | 13/13 | $\begin{gathered} 1.575 \pm 0.917 \\ (0.575-3.241) \end{gathered}$ |  |  |
| Flathead catfish | 8/8 | $\begin{gathered} 0.842 \pm 0.456 \\ (0.260-1.689) \\ \hline \end{gathered}$ |  |  |
| Hybrid striped bass | 2/2 | $\begin{gathered} \hline 3.803 \pm 3.005 \\ (1.678-5.928) \\ \hline \end{gathered}$ |  |  |
| Largemouth bass | 27/28 | $\begin{gathered} 0.556 \pm 0.432 \\ (N D-1.927) \\ \hline \end{gathered}$ |  |  |
| Spotted gar | 4/4 | $\begin{gathered} \hline 0.969 \pm 0.507 \\ (0.253-1.416) \\ \hline \end{gathered}$ | 3.49 |  |
| White bass | 2/2 | $\begin{gathered} \hline 0.907 \pm 0.195 \\ (0.769-1.045) \\ \hline \end{gathered}$ |  |  |
| White crappie | 3/3 | $\begin{gathered} \hline 0.169 \pm 0.109 \\ (0.091-0.294) \\ \hline \end{gathered}$ |  |  |
| Crappie Spp. (black and white) | 6/7 | $\begin{gathered} \hline 0.139 \pm 0.121 \\ \text { (ND-0.294) } \\ \hline \end{gathered}$ |  |  |
| Temperate basses | 4/4 | $\begin{gathered} \hline 2.355 \pm 2.412 \\ (0.769-5.928) \end{gathered}$ |  |  |
| All fish combined | 98/100 | $\begin{aligned} & 1.013 \pm 1.340 \\ & \text { (ND-10.400) } \end{aligned}$ |  |  |

[^13]Table 6.0. Volatile organic compounds ( $\mathrm{mg} / \mathrm{kg}$ ) in fish collected from the Ellison Creek Reservoir by species, 2014.

| Species | Number Detected/ Number Tested | Mean $\pm$ S.D. <br> (Min-Max) | HAC Value (nonca; mg/kg) | Basis for Comparison Value |
| :---: | :---: | :---: | :---: | :---: |
| Trichlorofluoromethane |  |  |  |  |
| Channel catfish | 8/8 | $\begin{gathered} 0.090 \pm 0.105 \\ (0.013-0.332) \end{gathered}$ |  |  |
| Largemouth bass | 7/8 | $\begin{gathered} \hline 0.032 \pm 0.026 \\ \text { (ND-0.079) } \\ \hline \end{gathered}$ | 700 | EPA Chronic Oral RfD - 3.0E-1 (mg/kg)/day |
| All fish combined | 15/16 | $\begin{aligned} & \hline 0.061 \pm 0.080 \\ & \text { (ND }-0.332 \text { ) } \\ & \hline \hline \end{aligned}$ |  |  |

Table 7. Hazard quotients (HQs) for mercury in fish collected from Ellison Creek Reservoir in 2014. Table 7 also provides suggested weekly eight-ounce meal consumption rates for $70-\mathrm{kg}$ adults. ${ }^{\text {r }}$

| Species | Number of Samples | Hazard Quotient | Meals per Week |
| :--- | :---: | :---: | :---: |
| Bowfin | 3 | 0.55 | 1.7 |
| Channel catfish | 35 | 0.05 | unrestricted $^{\text {s }}$ |
| Common carp | 13 | 0.06 | unrestricted |
| Flathead catfish | 8 | 0.09 | 10.4 |
| Hybrid striped bass | 2 | 0.24 | 3.9 |
| Largemouth bass | 30 | 0.18 | 5.2 |
| Spotted gar | 2 | 0.14 | 6.7 |
| White bass | 7 | 0.15 | 6.3 |
| Crappie Spp. (black and <br> white) | 4 | 0.03 | 15.8 |
| Sunfish Spp. (Bluegill, <br> Green, Redbreast) | 108 | 0.19 | unrestricted |
| Temperate basses | 4 | 0.11 | 4.8 |
| All fish combined |  | 8.3 |  |

[^14]Table 8.1. Hazard quotients (HQs) and hazard indices (HIs) for PCBs and/or PCDDs/PCDFs in fish collected from Ellison Creek Reservoir in 2014. Table 8.1 also provides suggested weekly eight-ounce meal consumption rates for 70-kg adults. ${ }^{\text {t }}$

| Contaminant/Species | Number of Samples | Hazard Quotient | Meals per Week |
| :---: | :---: | :---: | :---: |
| Bowfin |  |  |  |
| PCBs | 3 | 0.58 | 1.6 |
| PCDDs/PCDFs |  | 0.09 | 10.0 |
| Hazard Index (meals per week) |  | 0.67 | 1.4 |
| Channel catfish |  |  |  |
| PCBs | 35 | $2.38{ }^{\text {u }}$ | $0.4{ }^{\text {v }}$ |
| PCDDs/PCDFs | 33 | 0.57 | 1.6 |
| Hazard Index (meals per week) |  | 2.95 | 0.3 |
| Common carp |  |  |  |
| PCBs | 13 | 4.48 | 0.2 |
| PCDDs/PCDFs |  | 0.68 | 1.4 |
| Hazard Index (meals per week) |  | 5.15 | 0.2 |
| Flathead catfish |  |  |  |
| PCBs | 8 | 2.04 | 0.5 |
| PCDDs/PCDFs |  | 0.36 | 2.6 |
| Hazard Index (meals per week) |  | 2.40 | 0.4 |
| Hybrid striped bass |  |  |  |
| PCBs | 2 | 5.85 | 0.2 |
| PCDDs/PCDFs |  | 1.63 | 0.6 |
| Hazard Index (meals per week) |  | 7.48 | 0.1 |

[^15]Table 8.2. Hazard quotients (HQs) and hazard indices (HIs) for PCBs and/or PCDDs/PCDFs in fish collected from Ellison Creek Reservoir in 2014. Table 8.2 also provides suggested weekly eight-ounce meal consumption rates for $70-\mathrm{kg}$ adults. ${ }^{\text {w }}$

| Contaminant/Species | Number of Samples | Hazard Quotient | Meals per Week |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Largemouth bass | 30 | $1.37^{\mathrm{x}}$ | $\mathbf{0 . 7}^{\mathrm{y}}$ |  |  |
| PCBs | 28 | 0.24 | 3.9 |  |  |
| PCDDs/PCDFs | $\mathbf{1 . 6 1}$ | $\mathbf{0 . 6}$ |  |  |  |
| Hazard Index (meals per week) |  |  |  |  |  |


| Spotted gar |  |  |  |
| :--- | :--- | :--- | :--- |
| PCBs |  |  |  |
| PCDDs/PCDFs |  | $\mathbf{0 . 4}$ |  |
|  |  | 0.42 | 2.2 |
| Hazard Index (meals per week) |  | $\mathbf{2 . 7 9}$ | $\mathbf{0 . 3}$ |

White bass

| PCBs | 2 | 1.56 | 0.6 |
| :---: | :---: | :---: | :---: |
| PCDDs/PCDFs |  | 0.39 | 2.4 |
| Hazard Index (meals per week) |  | 1.95 | 0.5 |
| Crappie Spp. (black and white) |  |  |  |
| PCBs | 7 | 0.47 | 2.0 |
| PCDDs/PCDFs |  | 0.06 | 15.5 |
| Hazard Index (meals per week) |  | 0.53 | 1.7 |
| Sunfish Spp. (Bluegill, Green, Redbreast) |  |  |  |
| PCBs | 3 | 2.21 | 0.4 |
| Hazard Index (meals per week) |  | 2.21 | 0.4 |

[^16]Table 8.3. Hazard quotients (HQs) and hazard indices (HIs) for PCBs and/or PCDDs/PCDFs in fish collected from Ellison Creek Reservoir in 2014. Table 8.3 also provides suggested weekly eight-ounce meal consumption rates for $70-\mathrm{kg}$ adults. ${ }^{\text {² }}$

| Contaminant/Species | Number of Samples | Hazard Quotient | Meals per Week |  |
| :--- | :---: | :---: | :---: | :---: |
| Temperate basses |  |  |  |  |
| PCBs |  | $3.71^{\text {aa }}$ | $0.2^{\text {bb }}$ |  |
| PCDDs/PCDFs | 1.01 | 0.9 |  |  |
| Hazard Index (meals per week) |  | 4.72 | 0.2 |  |
| All fish combined | 107 | 2.19 | 0.4 |  |
| PCBs | 100 | 0.43 | 2.1 |  |
| PCDDs/PCDFs | 2.62 | 0.4 |  |  |
|  |  |  |  |  |

[^17]Table 9.1. Calculated theoretical lifetime excess cumulative cancer risk from consuming fish collected in 2014 from Ellison Creek Reservoir containing carcinogens and suggested consumption rate (eight-ounce meals/week) for 70 kg adults who regularly eat fish from Ellison Creek Reservoir over a 30-year period.cc

| Species/Contaminant | Number of Samples | Theoretical Lifetime Excess Cancer Risk |  | Meals per Week |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Risk | Population Size that Would Result in One Excess Cancer |  |
| Bowfin |  |  |  |  |
| PCBs | 3 | $9.9 \mathrm{E}-06$ | 100,823 | 9.3 |
| PCDDs/PCDFs |  | 6.2E-06 | 161,575 | 14.9 |
| Arsenic |  | 4.4E-06 | 226,852 | unrestricted ${ }^{\text {dd }}$ |
| Cumulative Cancer Risk |  | $2.1 \mathrm{E}-05$ | 48,743 | 4.5 |
| Channel catfish |  |  |  |  |
| Arsenic | 33 | $1.9 \mathrm{E}-06$ | 518,519 | unrestricted |
| Chlordane | 8 | 7.07E-07 | 1,414,141 | unrestricted |
| DDE |  | $1.79 \mathrm{E}-05$ | 55,783 | 5.2 |
| PCBs | 35 | $4.1 \mathrm{E}-05$ | 24,525 | 2.3 |
| PCDDs/PCDFs | 33 | 3.8E-05 | 26,380 | 2.4 |
| Cumulative Cancer Risk |  | $9.9 \mathrm{E}-05^{\text {ee }}$ | 10,076 | $0.9{ }^{\text {ff }}$ |
| Common carp |  |  |  |  |
| Arsenic | 13 | $1.4 \mathrm{E}-06$ | 725,926 | unrestricted |
| PCBs |  | 7.7E-05 | 13,025 | 1.2 |
| PCDDs/PCDFs |  | $4.5 \mathrm{E}-05$ | 22,159 | 2.0 |
| Cumulative Cancer Risk |  | $1.2 \mathrm{E}-04$ | 8,112 | 0.7 |

[^18]| Species/Contaminant | Number of Samples | Theoretical Lifetime Excess Cancer Risk |  | Meals per Week |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Risk | Population Size that Would Result in One Excess Cancer |  |
| Flathead catfish |  |  |  |  |
| Arsenic | 8 | 1.9E-06 | 518,519 | unrestricted ${ }^{\text {hh }}$ |
| PCBs |  | 3.5E-05 | 28,655 | 2.6 |
| PCDDs/PCDFs |  | 2.4E-05 | 41,449 | 3.8 |
| Cumulative Cancer Risk |  | 6.1E-05 | 16,406 | 1.5 |
| Hybrid striped bass |  |  |  |  |
| PCBs | 2 | $1.0 \mathrm{E}-04^{\text {ii }}$ | 9,972 | $0.9{ }^{\text {jij }}$ |
| PCDDs/PCDFs |  | 1.1E-04 | 9,177 | 0.8 |
| Cumulative Cancer Risk |  | 2.1E-04 | 4,779 | 0.4 |
| Largemouth bass |  |  |  |  |
| Chlordane | 8 | 1.67E-07 | 5,982,906 | unrestricted |
| DDE |  | 5.88E-06 | 170,139 | 15.7 |
| PCBs | 30 | 2.4E-05 | 42,535 | 3.9 |
| PCDDs/PCDFs | 28 | 1.6E-05 | 62,770 | 5.8 |
| Cumulative Cancer Risk |  | 4.7E-05 | 21,091 | 1.9 |
| Spotted gar |  |  |  |  |
| PCBs | 4 | 4.1E-05 | 24,525 | 2.3 |
| PCDDs/PCDFs |  | 2.8E-05 | 36,017 | 3.3 |
| Cumulative Cancer Risk |  | 6.9E-05 | 14,590 | 1.3 |

[^19]| Species/Contaminant | Number of Samples | Theoretical Lifetime Excess Cancer Risk |  | Meals per Week |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |
|  |  | Risk | Population Size that Would Result in One Excess Cancer |  |
| White bass |  |  |  |  |
| Arsenic | 2 | 6.1E-06 | 164,983 | 15.2 |
| PCBs |  | 2.7E-05 | 37,291 | 3.4 |
| PCDDs/PCDFs |  | 2.6E-05 | 38,479 | 3.6 |
| Cumulative Cancer Risk |  | 5.9E-05 | 16,988 | 1.6 |
| Crappie Spp. (black and white) |  |  |  |  |
| Arsenic | 7 | $2.8 \mathrm{E}-06$ | 362,963 | unrestricted" |
| PCBs |  | $8.1 \mathrm{E}-06$ | 123,737 | 11.4 |
| PCDDs/PCDFs |  | 4.0E-06 | 251,081 | unrestricted |
| Cumulative Cancer Risk |  | 1.5E-05 | 67,479 | 6.2 |
| Sunfish Spp. (Bluegill, Green, Redbreast) |  |  |  |  |
| PCBs | 3 | $3.8 \mathrm{E}-05$ | 26,429 | 2.4 |
| Cumulative Cancer Risk |  | $3.8 \mathrm{E}-05$ | 26,429 | 2.4 |

[^20]| Species/Contaminant | Number of Samples | Theoretical Lifetime Excess Cancer Risk |  | Meals per Week |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |
|  |  | Risk | Population Size that Would Result in One Excess Cancer |  |
| Temperate basses |  |  |  |  |
| Arsenic | 4 | 3.6E-06 | 279,202 | unrestricted ${ }^{\text {nn }}$ |
| PCBs |  | 6.4E-05 | 15,735 | 1.5 |
| PCDDs/PCDFs |  | 6.7E-05 | 14,820 | 1.4 |
| Cumulative Cancer Risk |  | $1.3 \mathrm{E}-04{ }^{\circ}$ | 7,429 | $0.7{ }^{\text {pp }}$ |
| All fish combined |  |  |  |  |
| Arsenic | 100 | 1.9E-06 | 518,519 | unrestricted |
| Chlordane | 16 | 4.37E-07 | 2,287,582 | unrestricted |
| DDE |  | $1.18 \mathrm{E}-05$ | 85,069 | 7.9 |
| PCBs | 107 | $3.7 \mathrm{E}-05$ | 26,668 | 2.5 |
| PCDDs/PCDFs | 100 | $2.9 \mathrm{E}-05$ | 34,452 | 3.2 |
| Cumulative Cancer Risk |  | 8.1E-05 | 12,405 | 1.1 |

[^21]Table 10. SALG recommended fish consumption advice for Ellison Creek Reservoir, 2014.

| Contaminants of Concern | Species | Women of childbearing age and children < $12^{\text {q9 }}$ | Women past childbearing age and males 12 and older ${ }^{\text {rr }}$ |
| :---: | :---: | :---: | :---: |
| Dioxins and PCBs | Channel catfish | DO NOT EAT | 1 meal/month |
|  | Common carp | DO NOT EAT | DO NOT EAT |
|  | Flathead catfish | DO NOT EAT | 1 meal/month |
|  | Hybrid striped bass | DO NOT EAT | DO NOT EAT |
|  | Largemouth bass | 1 meal/month | 2 meals/month |
|  | Spotted gar | DO NOT EAT | 1 meal/month |
|  | Sunfish Spp. (Bluegill, Green, Redbreast) | DO NOT EAT | 1 meal/month |
|  | White bass | DO NOT EAT | 2 meals/month |

[^22]
## LITERATURE CITED

${ }^{1}$ Texas Commission on Environmental Quality (TCEQ). 2004 Texas water quality inventory and 303(d) list. Cypress Creek Basin, segment 0404A Ellison Creek Reservoir fact sheet. https://www.tceq.texas.gov/assets/public/compliance/monops/water/assessments/04 0404A fact.pdf (Accessed April 22, 2015).
${ }^{2}$ Texas Department of State Health Services (DSHS). 2005. Characterization of potential health effects associated with consumption of fish from Ellison Creek Reservoir, Morris County, Texas. http://www.dshs.state.tx.us/seafood/PDF2/Risk-Characterization/EllisonCreekReservoirRC2005RevisedFebruary2007.pdf (Accessed April 22, 2015).
${ }^{3}$ Texas Department of State Health Services (DSHS). 2005. Fish and shellfish consumption advisory 29 (ADV-29) Ellison Creek Reservoir, November 28, 2005. http://www.dshs.state.tx.us/seafood/advisories-bans.aspx (Accessed April 5, 2016).
${ }^{4}$ Clean Water Act. 33 USC 125 et seq. 40CFR part 131: Water Quality Standards.
${ }^{5}$ Texas State Soil and Water Conservation Board (TSSWCB), Total Maximum Daily Load Program. https://www.tsswcb.texas.gov/en/tmdl (Accessed May 28, 2015).
${ }^{6}$ Texas Commission on Environmental Quality (TCEQ). Texas integrated report of surface water quality. http://www.tceq.state.tx.us/waterquality/assessment (Accessed May 28, 2015).
${ }^{7}$ Bister, T.J. 2011. Inland fisheries division monitoring and management program 2010 survey report Lone Star Lake. Texas Parks and Wildlife Department, Federal Aid Project F-221-M-1, Texas Parks and Wildlife Department, Austin, Texas.
${ }^{8}$ Texas Parks and Wildlife Department (TPWD). Lone Star Lake, public access facilities. https://tpwd.texas.gov/fishboat/fish/recreational/lakes/lone star/access.phtml (Accessed April 22, 2015).
${ }^{9}$ United States Census Bureau (USCB). State and county quickfacts. http://quickfacts.census.gov/qfd/states/48/48343.html (Accessed April 22, 2015).
${ }^{10}$ United States Census Bureau (USCB). City quickfacts. http://quickfacts.census.gov/qfd/states/48/4843888.html (Accessed April 22, 2015).

11 United States Environmental Protection Agency (USEPA). 2004. Economic and benefits analysis for the proposed section 316(b) phase II existing facilities rule. http://water.epa.gov/lawsregs/lawsguidance/cwa/316b/upload/Cooling-Water Phase-2 Economics 2004.pdf (Accessed October 1, 2014).
${ }^{12}$ Texas Department of State Health Services (DSHS). 2007. Standard operating procedures and quality assurance/quality control manual. Seafood and Aquatic Life Group Survey Team, Austin, Texas.

13 United States Environmental Protection Agency (USEPA). 2000. Guidance for assessing chemical contaminant data for use in fish advisories. vol. 1, fish sampling and analysis, $3^{\text {rd }}$ ed. EPA-823-B-00-007. Office of Water, Washington, D.C.
${ }^{14}$ Toxic Substances Coordinating Committee (TSCC) Web site. http://www.tscc.state.tx.us/default.htm (Accessed November 20, 2014).
${ }^{15}$ Gulf States Marine Fisheries Commission (GSMFC). 2009. Practical handbook for determining the ages of Gulf of Mexico fishes, $2^{\text {nd }}$ Edition. GSMFC Publication Number 167. Ocean Springs, MS.
${ }^{16}$ Texas Parks and Wildlife Department (TPWD). 2009. Texas inland fishery assessment procedures, TPWD Inland Fisheries Division unpublished manual. Austin, TX.
${ }^{17}$ United States Environmental Protection Agency (USEPA). Polychlorinated biphenyls (PCBs). Learn about polychlorinated biphenyls. PCB congeners and homologs. https://www.epa.gov/pcbs/learn-about-polychlorinated-biphenyls-pcbs\#congeners (Accessed April 5, 2016).

18 United States Environmental Protection Agency (USEPA). 2000. Guidance for assessing chemical contaminant data for use in fish advisories. vol. 2, risk assessment and fish consumption limits, $3^{\text {rd }}$ ed. EPA-823-00-008. Office of Water, Washington, D.C.
${ }^{19}$ Agency for Toxic Substances and Disease Registry (ATSDR). 2007. Toxicological profile for arsenic. United States Department of Health \& Human Services, Public Health Service Atlanta, GA.
${ }^{20}$ Clean Water Act (CWA). 33 USC 125 et seq. 40CFR part 131: Water Quality Standards.
${ }^{21}$ Agency for Toxic Substances and Disease Registry (ATSDR). 1999. Toxicological profile for mercury (update). United States Department of Health \& Human Services, Public Health Service. Atlanta, GA.
${ }^{22}$ Geochemical and Environmental Research Group (GERG). 1998. Standard operating procedures (SOP-9727). Determination of percent lipid in biological tissue.
${ }^{23}$ United States Environmental Protection Agency (USEPA). Polychlorinated biphenyls (PCBs). Learn about polychlorinated biphenyls. PCB mixtures and trade names. https://www.epa.gov/pcbs/learn-about-polychlorinated-biphenyls-pcbs\#congeners (Accessed April 5, 2016).
${ }^{24}$ Lauenstein, G.G. \& Cantillo, A.Y. 1993. Sampling and analytical methods of the national status and trends program national benthic surveillance and mussel watch projects 1984-1992: overview and summary of methods Vol. I. NOAA Tech. Memo 71. NOAA/CMBAD/ORCA. Silver Spring, MD.
157pp. http://docs.lib.noaa.gov/noaa documents/NOS/ORCA/TM NOS ORCA/nos orca 71v1.pdf (Accessed April 5, 2016).
${ }^{25}$ McFarland, V.A. \& Clarke, J.U. 1989. Environmental occurrence, abundance, and potential toxicity of polychlorinated biphenyl congeners: considerations for a congener-specific analysis. Environmental Health Perspectives. 81:225-239.
${ }^{26}$ Integrated Risk Information System (IRIS). Polychlorinated biphenyls (PCBs) (CASRN 1336-36-3), Part II, B.3. United States Environmental Protection Agency. http://www.epa.gov/iris/subst/0294.htm (Accessed November 20, 2014).
${ }^{27}$ Integrated Risk Information System (IRIS). Comparison of database information for RfDs on Aroclor ${ }^{*}$ 1016, 1254, 1260. United States Environmental Protection Agency. http://cfpub.epa.gov/ncea/iris/compare.cfm (Accessed November 20, 2014).
${ }^{28}$ Van den Berg, M., L. Birnbaum, ATC Bosveld et al. 1998. Toxic equivalency factors (TEFs) for PCBs, PCDDs, PCDFs for humans and wildlife. Environ. Health Perspect. 106(12):775-792.
${ }^{29}$ World Health Organization (WHO). 2005. Project for the re-evaluation of human and mammalian toxic equivalency factors (TEFs) of dioxins and dioxin-like compounds. http://www.who.int/ipcs/assessment/public health/dioxins other/en/ (Accessed November 20, 2014).
${ }^{30}$ De Rosa, CT, D. Brown, R. Dhara et al. 1997. Dioxin and dioxin-like compounds in soil, part 2: Technical support document for ATSDR interim policy guidline. Toxicol. Ind. Health. 13(6):759-768. http://www.atsdr.cdc.gov/hac/pha/midlandsoil-hc060304/appendixesept1.pdf (Accessed November 20, 2014).
${ }^{31}$ Klaassen C.D., editor. 2001. Casarett and Doull's toxicology: the basic science of poisons, $6^{\text {th }}$ ed. McGraw-Hill Medical Publishing Division, New York, NY.
${ }^{32}$ Integrated Risk Information System (IRIS). 1993. Reference dose (RfD): description and use in risk assessments. United States Environmental Protection Agency. http://www.epa.gov/iris/rfd.htm (Accessed November 24, 2014).
${ }^{33}$ Agency for Toxic Substances and Disease Registry (ATSDR). 2009. Minimal risk levels for hazardous substances. United States Department of Health \& Human Services. Public Health Service. http://www.atsdr.cdc.gov/mrls/index.html (Accessed November 24, 2014).
${ }^{34}$ Integrated Risk Information System (IRIS). 2010. IRIS glossary/acronyms \& abbreviations. United States Environmental Protection Agency. http://ofmpub.epa.gov/sor internet/registry/termreg/searchandretrieve/glossariesandkeywordlists/search.do?de tails=\&glossaryName=IRIS\%20Glossary (Accessed November 24, 2014).
${ }^{35}$ United States Environmental Protection Agency (USEPA). 1999. Glossary of key terms. Technology transfer network national-scale air toxics assessment. https://www.epa.gov/national-air-toxics-assessment (Accessed November 24, 2014).
${ }^{36}$ Thompson, K.M. 2004. Changes in children's exposure as a function of age and the relevance of age definitions for exposure and health risk assessment. MedGenMed. 6(3), 2004.
http://www.medscape.com/viewarticle/480733. (Accessed November 24, 2014).
${ }^{37}$ University of Minnesota, Maternal and Child Health Program, School of Public Health. 2004. Children's special vulnerability to environmental health risks. Healthy Generations 4(3). http://www.epi.umn.edu/mch/wpcontent/uploads/2015/09/hg february04.pdf (Accessed November 24, 2014). Page has moved
${ }^{38}$ Selevan, S.G., C.A. Kimmel, and P. Mendola. 2000. Identifying critical windows of exposure for children's health. Environmental Health Perspectives Volume 108, Supplement 3.
${ }^{39}$ Schmidt, C.W. 2003. Adjusting for youth: updated cancer risk guidelines. Environmental Health Perspectives. 111(13): A708-A710.
${ }^{40}$ Agency for Toxic Substances and Disease Registry (ATSDR). 1995. Child health initiative. United States Department of Health \& Human Services. Public Health Service. ATSDR Office of Children's Health. Atlanta, GA.
${ }^{41}$ United States Environmental Protection Agency (USEPA). 2000. Strategy for research on environmental risks to children, Section 1 and 2. Office of Research and Development (ORD) Washington, D.C.
${ }^{42}$ Systat 13 for Windows ${ }^{\circledR}$. Version 13.1. Copyright® Systat Software, Inc., 2009 all rights reserved. http://www.systat.com/ (Accessed November 24, 2014).
${ }^{43}$ Microsoft Corporation. Microsoft ${ }^{\circledR}$ Office Excel 2003. Copyright ${ }^{\oplus}$ Microsoft Corporation 1985-2003.
${ }^{44}$ Centers for Disease Control and Prevention (CDC). 2005. Preventing lead poisoning in young children. United States Department of Health \& Human Services. Atlanta, GA.
http://www.cdc.gov/nceh/lead/publications/PrevLeadPoisoning.pdf (November 24, 2014).
${ }^{45}$ Centers for Disease Control and Prevention (CDC). 2007. Interpreting and managing blood lead levels <10 $\mathrm{mcg} / \mathrm{dL}$ in children and reducing childhood exposures to lead. United States Department of Health \& Human Services, CDC Advisory Committee on Childhood Lead Poisoning Prevention. Atlanta, GA. MMWR 56(RR08); 1-14; 16. http://www.cdc.gov/mmwr/preview/mmwrhtml/rr5608a1.htm (Accessed November 24, 2014). ERRATUM MMWR November 30, 2007 / 56(47):1241-1242. http://www.cdc.gov/mmwr/preview/mmwrhtml/mm5647a4.htm (Accessed November 24, 2014).
${ }^{46}$ Texas Parks and Wildlife Department (TPWD). 2015. Outdoor annual hunting and fishing regulations. http://tpwd.texas.gov/publications/pwdpubs/media/cs bk k0700 0284.pdf (valid September 1, 2015 through August 31, 2016; Accessed February 23, 2016).
${ }^{47}$ United States Environmental Protection Agency (USEPA). 1996. Guidance for assessing chemical contaminant data for use in fish advisories. vol. 3, overview of risk management. EPA-823-B-96-006. Office of Water, Washington, D.C.
${ }^{48}$ Texas Statutes: Health and Safety Code, Chapter 436, Subchapter D, §436.061and § 436.091.
${ }^{49}$ Department of State Health Services (DSHS). 2009. Guide to eating Texas fish and Crabs. Seafood and Aquatic Life Group. Austin, TX.
${ }^{50}$ Department of State Health Services (DSHS). 2014. Seafood and Aquatic Life Group Web site. Austin, TX. http://www.dshs.state.tx.us/seafood/ (Accessed November 24, 2014).
${ }^{51}$ Texas Department of Health (DSHS). 2003. Quantitative risk characterization Brandy Branch Reservoir. Seafood Safety Division. Austin, TX.


[^0]:    ${ }^{\text {a }}$ The terms DSHS and SALG may be used interchangeably throughout this document and mean the same agency.

[^1]:    ${ }^{b}$ A PCB congener is any single, unique well-defined chemical compound in the PCB category. The name of a congener specifies the total number of chlorine substituents and the position of each chlorine (e.g., 4,4' dichlorobiphenyl is a congener comprising the biphenyl structure with two chlorine substituents, one on each of the number 4 carbons of the two rings). In 1980, a numbering system was developed, which assigned a sequential number to each of the 209 PCB congeners.

[^2]:    ${ }^{\text {c }}$ PCB homologs are subcategories of PCB congeners having equal numbers of chlorine substituents (e.g., the tetrachlorobiphenyls are all PCB congeners with exactly four chlorine substituents that may be in any arrangement).
    ${ }^{d}$ Aroclor is a PCB mixture produced from 1930 to 1979. It is one of the most commonly known trade names for PCB mixtures. There are many types of Aroclors and each has a distinguishing suffix number that indicates the degree of chlorination. The numbering standard is as follows: The first two digits refer to the number of carbon atoms in the phenyl rings and the third and fourth digits indicate the percentage of chlorine by mass in the mixture (e.g., Aroclor 1254 means that the mixture has 12 carbon atoms and contains $54 \%$ chlorine by weight).

[^3]:    ${ }^{e}$ A tolerable upper intake level (UL) is the highest average daily nutrient intake level that is likely to pose no risk of adverse health effects to almost all individuals in the general population. As intake increases above the UL, the potential risk of adverse effects may increase. The UL represents total intake from food, water, and supplements.

[^4]:    ${ }^{f}$ "J-value" is standard laboratory nomenclature for analyte concentrations that are detected and reported below the reporting limit (<RL). The reported concentration is considered an estimate, quantitation of which may be suspect and may not be reproducible. The DSHS treats J-Values as "not detected" in its statistical analyses of a sample set.
    ${ }^{\mathrm{g}}$ The SALG risk assessors' rationale for computing PCDDs/PCDFs descriptive statistics using the aforementioned method is based on the proximity of the laboratory reporting limits and the health assessment comparison value for PCDDs/PCDFs. Thus, applying the standard SALG method utilizing $1 / 2$ the reporting limit for analytes designated as not detected (ND) or estimated (J) will likely overestimate the PCDDs/PCDFs fish tissue concentration.

[^5]:    ${ }^{\mathrm{h}}$ Most arsenic in fish and shellfish occurs as organic arsenic, considered virtually nontoxic. For risk assessment calculations, DSHS assumes that total arsenic is composed of $10 \%$ inorganic arsenic in fish and shellfish tissues. ${ }^{i}$ Derived from the MRL or RfD for noncarcinogens or the EPA slope factor for carcinogens; assumes a body weight of 70 kg , and a consumption rate of 30 grams per day, and assumes a 30 -year exposure period for carcinogens and an excess lifetime cancer risk of $1 \times 10^{-4}$.

[^6]:    ${ }^{j}$ The Food and Nutrition Board, Institute of Medicine, National Academies UL for copper is $10 \mathrm{mg} /$ day.

[^7]:    ${ }^{k}$ The DSHS applied relative source contribution methodology (RSC) developed by EPA to derive a HAC value for selenium. DSHS risk assessor's assumed that $50 \%$ of the daily selenium intake is from other foods or supplements ( $\approx 200 \mu \mathrm{~g} / \mathrm{day}$ for a 70 kg adult or one-half the RfD) and subtracted an amount equal to $50 \%$ of the RfD from the RfD to account for other sources of exposure to selenium. The remainder of the RfD, $0.0025 \mathrm{mg} / \mathrm{kg} / \mathrm{day}$, was utilized to calculate the HAC value for selenium.

[^8]:    ${ }^{1}$ Emboldened numbers denote that PCB concentrations equal and/or exceed the DSHS HAC value for PCBs.

[^9]:    ${ }^{\mathrm{m}}$ Emboldened numbers denote that PCB concentrations equal and/or exceed the DSHS HAC value for PCBs.

[^10]:    ${ }^{\mathrm{n}}$ Emboldened numbers denote that PCB concentrations equal and/or exceed the DSHS HAC value for PCBs.

[^11]:    ${ }^{\circ}$ Emboldened numbers denote that PCDD/PCDF TEQ concentrations equal and/or exceed the DSHS HAC value for PCDDs/PCDFs.

[^12]:    ${ }^{p}$ Emboldened numbers denote that PCDD/PCDF TEQ concentrations equal and/or exceed the DSHS HAC value for PCDDs/PCDFs.

[^13]:    ${ }^{\text {q }}$ Emboldened numbers denote that PCDD/PCDF TEQ concentrations equal and/or exceed the DSHS HAC value for PCDDs/PCDFs.

[^14]:    ${ }^{r}$ DSHS assumes that children under 12 years of age and/or those that weigh less than 35 kg eat four-ounce meals.
    ${ }^{\mathrm{s}}$ Denotes that the allowable eight-ounce meals per week are > 16.0.

[^15]:    ${ }^{\mathrm{t}}$ DSHS assumes that children under 12 years of age and/or those that weigh less than 35 kg eat four-ounce meals.
    ${ }^{4}$ Emboldened numbers denote that the HQ or HI is $\geq 1.0$.
    ${ }^{\vee}$ Emboldened numbers denote that the calculated allowable meals for an adult are $\leq$ one meal per week.

[^16]:    ${ }^{\text {w }}$ DSHS assumes that children under 12 years of age and/or those that weigh less than 35 kg eat four-ounce meals.
    $\times$ Emboldened numbers denote that the HQ or HI is $\geq 1.0$.
    y Emboldened numbers denote that the calculated allowable meals for an adult are $\leq$ one meal per week.

[^17]:    ${ }^{2}$ DSHS assumes that children under 12 years of age and/or those that weigh less than 35 kg eat four-ounce meals.
    ${ }^{\text {aa }}$ Emboldened numbers denote that the HQ or HI is $\geq 1.0$.
    ${ }^{\text {bb }}$ Emboldened numbers denote that the calculated allowable meals for an adult are $\leq$ one meal per week.

[^18]:    cc DSHS assumes that children under 12 years of age and/or those who weigh less than 35 kg eat 4 -ounce meals.
    ${ }^{\text {dd }}$ Denotes that the allowable eight-ounce meals per week are > 16.0.
    ee Emboldened numbers denote calculated excess lifetime cancer risk after 30 years exposure is greater than 1.0E-04.
    ${ }^{\text {ff }}$ Emboldened numbers denote that the calculated allowable meals for an adult are $\leq$ one meal per week.

[^19]:    ${ }^{\text {gs }}$ DSHS assumes that children under 12 years of age and/or those who weigh less than 35 kg eat 4 -ounce meals.
    ${ }^{\text {hh }}$ Denotes that the allowable eight-ounce meals per week are > 16.0.
    ${ }^{\text {ii }}$ Emboldened numbers denote calculated excess lifetime cancer risk after 30 years exposure is greater than 1.0E-04.
    ij Emboldened numbers denote that the calculated allowable meals for an adult are $\leq$ one meal per week.

[^20]:    ${ }^{k k}$ DSHS assumes that children under 12 years of age and/or those who weigh less than 35 kg eat 4 -ounce meals.
    ${ }^{11}$ Denotes that the allowable eight-ounce meals per week are > 16.0.

[^21]:    mm DSHS assumes that children under 12 years of age and/or those who weigh less than 35 kg eat 4-ounce meals.
    ${ }^{n \mathrm{n}}$ Denotes that the allowable eight-ounce meals per week are > 16.0.
    ${ }^{\circ 0}$ Emboldened numbers denote calculated excess lifetime cancer risk after 30 years exposure is greater than 1.0E-04.
    ${ }^{\text {pp }}$ Emboldened numbers denote that the calculated allowable meals for an adult are $\leq$ one meal per week.

[^22]:    ${ }^{99}$ A meal is four ounces of fish.
    ${ }^{r r}$ A meal is eight ounces of fish.

